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**What level of native beetle diversity can be supported by forestry plantations? A global synthesis**

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1 **What level of native beetle diversity can be supported by forestry plantations? A global**  
2 **synthesis**

3

4 Running title: Forestry plantations and beetle diversity

5

6 **Abstract.** 1. Forestry plantations have been established globally to meet timber demands,  
7 often leading to the conversion of natural to artificial forests. Forestry plantations may support  
8 natural elements of forest biodiversity, but understanding their role in the maintenance of  
9 biodiversity is a crucial question.

10 2. We perform a meta-analysis of 48 studies to determine how forestry plantations relative to  
11 natural forests influence the species richness and abundance of three important coleopteran  
12 groups (*i.e.*, ground beetles, rove beetles, and dung beetles) of natural forests, given their  
13 essential role in ecosystem functioning.

14 3. We assessed whether beetle responses depended on taxonomic group, geographical  
15 location, native or exotic character of the planted tree species, and associated management  
16 characteristics (*i.e.*, composition, size, age, and connectivity of the plantations).

17 4. We found that forestry plantations negatively affected coleopteran species richness and  
18 abundance compared to natural forests. The negative impact was most severe in plantations  
19 with exotic tree species and located in tropical biomes.

20 5. Species richness and abundance of beetles significantly increased with plantation age in  
21 native plantations, but decreased in exotic ones. Also, small plantations close to native forest

22 had higher beetle species richness and abundance than ones located far away from native  
23 forest.

24 6. Stop the conversion of natural forests to plantations, the use of native tree species, and  
25 lengthening rotations is critical for allowing biodiversity recovery in forestry plantations,  
26 combined with a robust conservation strategy to protect threatened biodiversity and  
27 ecosystem functioning.

28

29 **Keywords:** Carabidae, Conservation, Ecological indicators, Exotic, Insect diversity, Native,  
30 Natural forest cover, Staphylinidae, Scarabaeidae.

31

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53

## 54 **Introduction**

55 Establishing forestry plantations has become a global strategy to enhance timber supply, and  
56 paper pulp demand (ITTO, 2016; Pirard *et al.*, 2016), and provide essential ecosystem  
57 services, especially climate change-mitigating carbon sequestration (Paul *et al.*, 2008) and  
58 watershed protection (Lamb *et al.*, 2005; Paquette & Messier, 2010; Baral *et al.*, 2016).

59 Governments have promoted forestry plantations through international agreements (*e.g.*, Bonn  
60 Challenge) and local or regional policies (Bull *et al.*, 2006), an action that has led to a  
61 worldwide increase of ca. 3 million ha annually between 2010 and 2020 (FAO, 2020).

62 Forestry plantations currently occupy ~290 million ha, of which ~60% is distributed

63 throughout the non-tropical biomes of Asia, Europe, and North America (FAO, 2020). In  
64 these areas, both native and exotic species (*e.g.*, *Eucalyptus* spp., *Fagus* spp., *Picea* spp.,  
65 *Pinus* spp., *Quercus* spp.) are extensively used. However, in tropical biomes, exotic species  
66 such as *Eucalyptus* spp. are used more frequently due to their fast growth and tolerance of a  
67 wide range of conditions (Simonetti *et al.*, 2012; Payn *et al.*, 2015; FAO, 2020).

68         Given the marked reduction in natural forests and their continued replacement by  
69 forestry plantations (see Keenan *et al.*, 2015; Payn *et al.*, 2015), it is crucial to understand the  
70 potential role of plantations as alternative habitats for biodiversity (Brockerhoff *et al.*, 2008).  
71 Forestry plantations frequently have a negative effect on vertebrate diversity (Barlow *et al.*,  
72 2007a), including amphibians (Kudavidanage *et al.*, 2012), birds (Castaño-Villa *et al.*, 2019),  
73 and mammals (Begotti *et al.*, 2018), or invertebrates, such as coleopterans (Horák *et al.*, 2019;  
74 Méndez-Rojas *et al.*, 2021) or butterflies (Kudavidanage *et al.*, 2012). However, other studies  
75 found biodiversity levels within plantations to match those found in natural forests,  
76 particularly non-tropical biomes (Plexida *et al.*, 2014; Magura *et al.*, 2015). A global  
77 understanding of the relative merits of forestry plantations as biodiversity supporters thus  
78 remains a core question. Ascertaining the characteristics associated with forestry plantations  
79 that favor higher biodiversity levels is important in making management recommendations.  
80 Studies have focused on the identity, origin (native or exotic) (Proença *et al.*, 2010; Horák *et*  
81 *al.*, 2019), composition of planted tree species (Piotto, 2008), as well as tree plantation  
82 purposes (*i.e.*, commercial or protective) (Brockerhoff *et al.*, 2008), planted area, age  
83 (Humphrey *et al.*, 1999; Kerr, 1999; Castaño-Villa *et al.*, 2019) and degree of isolation from  
84 natural forest remnants (Begotti *et al.*, 2018; Castaño-Villa *et al.*, 2019).

85         Coleoptera (Arthropoda: Insecta) form an exceptionally diverse group, distributed  
86 worldwide, that has important functional roles in almost every ecosystem (Samways, 2005;

87 Schowalter, 2008; New, 2010). In particular, ground beetles (Carabidae), rove beetles  
88 (Staphylinidae), and dung beetles (Scarabaeidae) contribute to several important ecosystem  
89 processes, including pest regulation, removal of organic matter, secondary seed dispersion,  
90 maintaining soil structure, and nutrient cycling (Rainio & Niemelä, 2003; Henle *et al.*, 2004;  
91 Spector, 2006; Nichols *et al.*, 2008; Magura *et al.*, 2015). They have been frequently used to  
92 quantify impacts associated with loss or degradation of natural forests (Nichols *et al.*, 2007;  
93 Magura *et al.*, 2017) and evaluate the effect of subsequent land-use changes (Rainio &  
94 Niemelä, 2003; Edwards *et al.*, 2014, 2017). Local studies on forestry plantations, when  
95 compared to natural forests, have shown a variety of responses in these coleoptera groups,  
96 from severe reductions to positive responses on species richness or abundance (see Gardner *et*  
97 *al.*, 2008; Taboada *et al.*, 2010; Gries *et al.*, 2012; Lange *et al.*, 2014; Magura *et al.*, 2015;  
98 Milheiras *et al.*, 2020). This indicates that specific management characteristics or geographic  
99 locations **modify** the effects of forestry plantations. We currently lack a global synthesis of  
100 beetle responses to plantation context and management, with previous evaluations restricted  
101 to single regions (*e.g.*, Palearctic; Paillet *et al.*, 2010), or not assessing the impacts of  
102 management characteristics (*e.g.*, Nichols *et al.*, 2007; Fuzessy *et al.*, 2021; Méndez-Rojas *et*  
103 *al.*, 2021).

104 Here, we perform a global meta-analysis on the response of ground, rove, and dung  
105 beetles to forestry plantations, including how their location and management characteristics  
106 (*e.g.*, age, size, origin, and a mixture of planted tree species) influence the diversity and  
107 abundance of these important reference groups. **Forestry plantations simplify forest**  
108 **composition and structure, and certain number of beetle species are closely linked with**  
109 **conditions that developed over long time with natural forest, as a result we predict that**  
110 forestry plantations will (H1) support lower species richness and abundance of the three

111 taxonomic groups than natural forests, and (H2) have positive effects modulated by native  
112 species implementation, or forestry plantations proximity to natural forests.

113

## 114 **Material and Methods**

### 115 *Literature search and inclusion criteria*

116 The literature search and selection followed the PRISMA methodology (Moher *et al.*, 2009),  
117 which only considers indexed articles (Nakagawa *et al.*, 2017). We searched the Scopus and  
118 Web of Science databases using the search terms: (forest\* OR regenerat\* OR plantat\* OR  
119 restorat\* OR “land-use”) AND (scarabaei\* OR “dung beetle\*” OR carabid\* OR “ground  
120 beetle\*” OR staphylinid\* OR “rove beetle\*”). We use these search terms based on the  
121 possibilities of inclusion of different articles assess a broad spectrum of land cover or land  
122 uses (including forestry plantations). We found 3675 articles, published between January  
123 1950 and July 2020. We deleted duplicate results, articles that include other families, or  
124 developed other topics, such as molecular biology, or behavioral studies. We retained 487  
125 studies containing the topic of interest, namely Carabidae, Staphylinidae, and/or Scarabaeidae  
126 species richness and/or abundance. The full text of these articles was assessed and filtered  
127 using the following criteria: (i) compared beetle species richness and/or abundance between  
128 natural forests (control) and forestry plantations (treatment), and (ii) presented descriptive  
129 sample statistics, including the mean values for species richness and/or abundance in control  
130 and treatment stands, their standard deviation (or other data from which these could be  
131 calculated), and sample sizes. Under these criteria, 48 articles were retained for data  
132 extraction (many of these articles sampled the richness or abundance of these beetle’s families  
133 by using pitfall traps; see Fig. S1).

134



135 (Fig. 1 here)

136

137 We used natural forest definition proposed by Chazdon *et al.*, (2016); and forestry  
138 plantations were classified as for commercial timber production or protective/conservation  
139 purposes (Stephens & Wagner, 2007; Simonetti *et al.*, 2012). Agricultural plantations (*e.g.*,  
140 cocoa, coffee, oil palm, rubber, or fruits; Castaño-Villa *et al.*, 2019) were excluded from this  
141 study. Some identified “restoration sites” (Audino *et al.*, 2014; Derhé *et al.*, 2016; Bowie *et*  
142 *al.*, 2019) were included, as tree planting and management are directly related to conservation  
143 efforts and classified as forestry plantations established for conservation purposes (Lamb *et*  
144 *al.*, 2005; Chazdon, 2008). When articles presented results as medians and interquartile  
145 ranges, we used the method proposed by Hozo *et al.* (2005) and modified by Wan *et al.*  
146 (2014) to obtain the means and standard deviations. Additionally, for articles that evaluated  
147 various sampling sites, geographical locations, or vegetation cover types (*e.g.*, Magura *et al.*,  
148 2002; Lange *et al.*, 2014) each case study was considered independently (Fontúrbel *et al.*,  
149 2015). Likewise, if articles presented different sampling periods (*e.g.*, several years), each  
150 was considered separately (Borenstein *et al.*, 2009; Mengersen *et al.*, 2013).

151

### 152 *Calculating the effect size*

153 Effect size was estimated using the Hedge’s *d* statistic, which uses weighted standardized  
154 deviations (Hedges & Olkin, 1985). The use of Hedge’s *d* is widespread in ecological studies  
155 that compare two groups since it presents the results in a continuous distribution (Gurevitch *et*  
156 *al.*, 2001; Borenstein *et al.*, 2009). We used random-effects models (Borenstein *et al.*, 2009)  
157 for which studies with different geographical locations (mostly primary studies) were the  
158 random effect, and plantation attributes (see below) were fixed effects. Random-effects

159 models also reduce the bias generated when comparing datasets of different sample sizes  
160 (Konstantopoulos & Hedges, 2010).

161

### 162 *Variables and moderators*

163 The global effect associated with the replacement of natural forests by forestry plantations  
164 was estimated collectively for species richness and abundance of the three selected beetle  
165 taxa. We analyzed the (fixed) effects of the following nine categorical moderator variables: (i)  
166 taxonomic group (*i.e.*, ground beetles, rove beetles, dung beetles), (ii) origin of the planted  
167 species (native or exotic), (iii) biome type (tropical between 23° N and 23° S or non-tropical),  
168 (iv) interaction between species origin and biome type (*i.e.*, non-tropical-exotic, non-tropical-  
169 native, tropical-exotic, tropical-native), (v) planted species (*i.e.*, *Eucalyptus*, *Quercus*, *Pinus*,  
170 and others), (vi) plantation composition (monoculture or mixed), (vii) plantation purpose  
171 (commercial or protective), (viii) plantation connectivity in the landscape (connected to or  
172 isolated from natural forests), and (ix) plantation size (small:  $\leq 400$  ha or large:  $\geq 1000$  ha;  
173 there were no plantations between these two size categories). To determine whether forestry  
174 plantations were connected to natural forests, we used the landscape definition of Driscoll *et*  
175 *al.* (2013), *i.e.*, connected plantations were defined as those in which the sampling location  
176 was separated from natural forest smaller than 500 m (Larsen & Forsyth, 2005; Hendrickx *et*  
177 *al.*, 2013; Cerda *et al.*, 2015; da Silva & Hernández, 2015).

178

### 179 *Statistical analysis*

180 To determine the heterogeneity between variables, we estimated the between-group  
181 homogeneity ( $Q_{\text{between}}$ ), a statistic with  $X^2$  distribution that compares the variation within and  
182 between the different levels of the variables (Higgins *et al.*, 2003). A  $Q_{\text{between}}$  value with  $p <$

183 0.05 indicates significant effect variation (*i.e.*, there is no unidirectional, common effect). We  
184 used this statistic because it is the best fit for the random-effects models we ran compared  
185 with other heterogeneity measurements such as  $I^2$  or  $\tau^2$  (Borenstein *et al.*, 2009). Finally, we  
186 constructed meta-regression models, separately for native and exotic plantations, using  
187 plantation age as the continuous variable to determine its effect on beetle species richness and  
188 abundance (Thompson & Higgins, 2002). Not all articles reported all variables, particularly  
189 tree species, plantation composition, age, size, or purpose. In this case, we only evaluated  
190 those from which information was available (Table S1).

191

#### 192 *Publication bias*

193 Meta-analyses may suffer from publication bias, resulting in missing studies (Borenstein *et*  
194 *al.*, 2009). As studies reporting neutral effects were unlikely published, we estimated the  
195 number of non-published neutral effect studies that would be necessary to obtain non-  
196 significant effects in our analyses using the Rosenthal test (Hillebrand, 2008). This test is  
197 robust when the confidence number is  $\geq 5N + 10$ , N being the number of case studies. We  
198 favored this test because of its reasonable adjustment to random model analyses (Jennions *et*  
199 *al.*, 2013). Duval & Tweedies' (2000) trim and fill method was used when significant  
200 asymmetry was found. This analysis evaluates the asymmetry in reported study outcomes and  
201 recalculates the average global effect and confidence intervals to validate whether the  
202 obtained results are reliable (Duval & Tweedie, 2000). Lastly, we used each article's ID as a  
203 random effect to determine whether the number of case studies per article influenced the  
204 direction and size of effects (see Supplementary results). All analyses were performed with  
205 the Comprehensive Meta-Analysis 3.0 software (Borenstein *et al.*, 2005).

206

## 207 **Results**

### 208 *Overview of forestry plantation database*

209 Of the 48 articles analyzed, we obtained 185 and 167 comparisons of beetle species richness  
210 and abundance, respectively, between plantations and natural forests (Table S1). Studies were  
211 distributed in 24 countries (Fig. 1): Brazil, China, and Poland had the most studies (we found  
212 articles performed in two countries or more; Table S2). Also, some boreal regions, especially  
213 North America (Canada and United States of North America), Scandinavian countries or  
214 Russia, did not present articles included in this meta-analysis. Mostly due to its exclusion due  
215 to the absence of the necessary statistical metrics (see inclusion criteria on methods). For the  
216 total number of comparisons, both non-tropical (68%) and tropical (22%) biomes were  
217 represented. Ground beetles were the most studied beetle group with 51% of the comparisons  
218 found, followed by dung beetles and rove beetles with 31.2% and 17.8%, respectively. Exotic  
219 plantations were studied in 49.44% of the comparisons, and most of them (84%) were planted  
220 as monocultures, often with *Pinus* spp. (23% of total comparisons came from such  
221 plantations) or *Eucalyptus* spp. (15% of total comparisons). Several mixed forestry  
222 plantations included exotic species such as *Acacia* or *Swietenia* spp. (25.9% of comparisons),  
223 and in some very particular cases abandoned plantations of exotic species were also used in  
224 forestry plantations established for conservation purposes (10% of comparisons).

225

### 226 *Overall beetle species richness and abundance*

227 Replacement of natural forests by forestry plantations was detrimental to overall species  
228 richness ( $d = -1.090$ ,  $CI = [-1.321, -0.859]$ ,  $p < 0.001$ ; Fig. 2a) and abundance ( $d = -0.438$ ,  $CI$   
229  $= [-0.646, -0.231]$ ,  $p < 0.001$ ; Fig. 3a) of beetles. Species richness did not vary among taxa

230 ( $Q_{\text{between}} = 2.568$ ,  $df = 2$ ,  $p = 0.276$ ; Fig. 2b). In contrast, beetle abundance differed among  
231 taxonomic groups ( $Q_{\text{between}} = 19.181$ ,  $df = 2$ ,  $p < 0.001$ ; Fig. 3b), with dung and rove beetles  
232 showing significantly more negative response than ground beetles.

233

234 (Fig. 2, 3 here)

235

### 236 *Effects of planted tree species and biome interaction*

237 Plantations of native species showed a smaller negative effect than plantations of exotic  
238 species either for species richness ( $Q_{\text{between}} = 26.729$ ,  $df = 1$ ,  $p < 0.001$ ; Fig. 2c) or abundance  
239 ( $Q_{\text{between}} = 5.852$ ,  $df = 1$ ,  $p = 0.015$ ; Fig. 3c). Biome type also affected species richness  
240 ( $Q_{\text{between}} = 11.971$ ,  $df = 1$ ,  $p < 0.001$ ; Fig. 2d) and abundance ( $Q_{\text{between}} = 17.088$ ,  $df = 1$ ,  $p <$   
241  $0.001$ ; Fig. 3d), with a greater negative effect of plantations in tropical biomes. Furthermore,  
242 the interaction between biome and the native or exotic character of plantations showed a  
243 significant effect for both species richness ( $Q_{\text{between}} = 67.132$ ,  $df = 3$ ,  $p < 0.001$ ) and  
244 abundance ( $Q_{\text{between}} = 26.934$ ,  $df = 3$ ,  $p < 0.001$ ).

245 Plantations composed of native species had negative effects only in the tropics, with  
246 no effect found in non-tropical biomes, while in exotic plantations detrimental effects were  
247 found in both non-tropical and tropical biomes (Fig. 2e; Fig 3e). Beetle species richness  
248 ( $Q_{\text{between}} = 149.66$ ,  $df = 9$ ,  $p < 0.001$ ; 2f) and abundance ( $Q_{\text{between}} = 38.189$ ,  $df = 9$ ,  $p < 0.001$ ;  
249 3f) was dependent on the genus of the planted tree species. Thus, plantations of *Acacia*,  
250 *Eucalyptus*, *Picea*, *Pinus*, and *Tectona* spp. caused significantly stronger negative effects than  
251 those of *Alnus*, *Eremanthus*, *Fagus*, or *Quercus* spp.

252

### 253 *Effects of plantation management*

254 The mixed vs. monocultural character of the plantations did not matter: both had a similarly  
255 negative effect on the species richness ( $Q_{\text{between}} = 1.139$ ,  $df = 1$ ,  $p = 0.285$ ; Fig. 2g) as well as  
256 the abundance ( $Q_{\text{between}} = 0.053$ ,  $df = 1$ ,  $P = 0.816$ ; Fig. 3g) of beetles. Likewise, the purpose  
257 (commercial or protective) for establishing a plantation had no effect on either beetle species  
258 richness ( $Q_{\text{between}} = 0.906$ ,  $df = 1$ ,  $p = 0.340$ ; Fig. 2h) or abundance ( $Q_{\text{between}} = 1.205$ ,  $df = 1$ ,  $p$   
259  $= 0.272$ ; Fig. 3h). However, forestry plantations establishing for protective purpose showed a  
260 neutral effect for beetle abundance.

261

#### 262 *Effects of the landscape around plantations*

263 We found a negative effect on species richness in forestry plantations that were isolated from  
264 natural forest, but a neutral response when a forestry plantation was connected to a natural  
265 forest ( $Q_{\text{between}} = 9.358$ ,  $df = 1$ ,  $p = 0.0022$ ; Fig. 2i). There was no impact of connectivity on  
266 abundance ( $Q_{\text{between}} = 0.973$ ,  $df = 1$ ,  $p = 0.323$ ; Fig. 3i), although we found a neutral effect for  
267 forestry plantations connected to natural forests. Species richness decreased significantly  
268 irrespective of plantation size ( $Q_{\text{between}} = 1.199$ ,  $df = 1$ ,  $p = 0.273$ ; Fig. 2j), whereas there was a  
269 size-dependent effect on abundance, which only decreased in large-sized plantations ( $Q_{\text{between}}$   
270  $= 7,598$ ,  $df = 1$ ,  $p = 0.0058$ ; Fig. 3j).

271

#### 272 *Effects of plantation age*

273 Plantation age had a significant effect on both beetle species richness and abundance (Fig. 4).  
274 However, the response direction was markedly different in plantations with native or exotic  
275 tree species; in native plantations, beetle species richness and abundance significantly  
276 increased with plantation age (slope = 0.0075,  $p < 0.001$ ;  $\text{Tau}^2 = 1.716$ ; Fig. 4a and slope =  
277 0.0065,  $p < 0.001$ ;  $\text{Tau}^2 = 2.125$ ; Fig. 4c; respectively), whereas in exotic plantations both

278 decreased significantly (slope = -0.0201,  $p < 0.001$ ;  $\text{Tau}^2 = 1.710$ ; Fig. 4b and slope = -0.017,  
279  $p < 0.001$ ;  $\text{Tau}^2 = 1.420$ ; Fig. 4d; respectively).

280

281 (Fig. 4 here)

282

### 283 *Publication bias*

284 The Rosenthal confidence test indicated that 9940 and 4213 case studies with neutral effects  
285 were necessary to obtain non-significant results for species richness and abundance in our  
286 analysis (when this value is compared with the safety threshold of 935 and 845 cases,  
287 respectively). The direction and size of the effect for species richness in ‘trim-and-fill’  
288 analysis did not vary; however, beetle abundance showed a neutral tendency (Table S3; Figs.  
289 S2, S3). We obtained significant heterogeneity values for species richness ( $Q_{\text{between}} = 767.023$ ,  
290  $df = 35$ ,  $p < 0.001$ ) and abundance ( $Q_{\text{between}} = 299.823$ ,  $df = 39$ ,  $p < 0.001$ ), indicating that the  
291 number of case studies per article did not distort the direction and size of effects. Our meta-  
292 analysis showed a small bias for beetle abundance due to the larger number of case studies  
293 with negative effects. However, our results should still be reliable and not affected by the  
294 omission of articles with neutral effects. Nevertheless, we recommend caution in interpreting  
295 the results associated with beetle abundance.

296

### 297 **Discussion**

298 [Our global meta-analysis on the effect of forestry plantations on native beetle diversity](#)  
299 [support our research hypotheses, revealing that species richness and abundance of ground,](#)  
300 [rove and dung beetles were generally lower than in natural forests \(H1\), and that geographical](#)

301 [location and management affected the conservation value of plantations \(H2\)](#). This points to  
302 the need to halt conversion of natural forest to plantation and the inclusion of native species  
303 on longer rotations.

304

305 *Forestry plantations support lower levels of beetle diversity than natural forests*

306 Our results showed [lower](#) species richness and abundance of beetle's assemblages in forestry  
307 plantations compared to natural forests. Previous single regional studies conclude that  
308 plantation expansion at the expense of natural forests contributes to global biodiversity loss,  
309 threatening native species assemblages, their functioning, and the ecosystem services they  
310 provide (Paillet *et al.*, 2010; Newbold *et al.*, 2015; Magura & Lövei, 2019). The negative  
311 response of the studied ground, rove, and dung beetle assemblages to forestry plantations also  
312 indicated the harmful consequences caused by [conversion of](#) natural ecosystems into  
313 [intensively managed lands](#) on organisms at various trophic levels (Barlow *et al.*, 2007a;  
314 Paillet *et al.*, 2010).

315 Ground and rove beetles are extremely sensitive to changes in environmental and  
316 habitat characteristics caused by plantation establishment (Pohl *et al.*, 2008; Koivula, 2011).  
317 Compared to natural forests, soil temperature in plantations is higher, and humidity is lower  
318 (Lange *et al.*, 2014; Senior *et al.*, 2017). Likewise, plantation maintenance results in a more  
319 open canopy and less leaf litter, coarse woody debris, fewer herbs, and shrubs compared to  
320 native forests (Paillet *et al.*, 2010; Lange *et al.*, 2014). These changes likely alter the  
321 availability of food: the density of other ground-dwelling invertebrates (for ground and rove  
322 beetles: Niemelä *et al.*, 2007; Magura *et al.*, 2015; Nagy *et al.*, 2015) and leaf-litter nutrients  
323 (for rove beetles: Barlow *et al.*, 2007b).



324 Multiple abiotic and biotic factors could explain changes in the abundance and  
325 diversity of dung beetle assemblages after the establishment of a plantation (Nichols *et al.*,  
326 2007). Habitat structure and microclimatic conditions (*i.e.*, light intensity, soil and air  
327 temperature, humidity), which differ in plantations versus **natural** forest, are crucial factors in  
328 determining the species composition (Hanski & Cambefort, 1991; Davis *et al.*, 2002; Gardner  
329 *et al.*, 2008) and community assembly (Audino *et al.*, 2017) of dung beetles. Moreover,  
330 forestry plantations can modify the availability of dung by influencing the presence/absence  
331 of mammals (Barlow *et al.*, 2007a) and, in turn, the quantity and quality of dung resources,  
332 causing a lower diversity of dung beetles (Gardner *et al.*, 2008).

333

334 *Plantation origin and biome are critical factors for beetle diversity*

335 The negative impact of plantations on beetle species richness and abundance was more  
336 pronounced in tropical than non-tropical biomes, supporting previous single regional studies  
337 (Grimbacher *et al.*, 2007; Chaudhary *et al.*, 2016) that emphasize that tropical species, are  
338 very sensitive to habitat alteration (**associated with prevalence of beetle forest specialists in**  
339 **tropical biomes due to the greater number of micro-habitats; see Halfpeter, 1991; Davis *et al.*,**  
340 **2001**). Our evaluation found that the implementation of exotic plantations had more  
341 detrimental effects on beetle species richness and abundance, **especially in the tropics and**  
342 **particularly related with tropical forest**, where such plantations are primarily composed of  
343 exotic species (Payn *et al.*, 2015; FAO, 2020). Local studies on invertebrate biodiversity in  
344 exotic forestry plantations revealed similar effects (Gardner *et al.*, 2008; Robson *et al.*, 2009;  
345 Roberge & Stenbacka, 2014; Nagy *et al.*, 2015; Beiroz *et al.*, 2016; Milheiras *et al.*, 2020).

346           The establishment of exotic tree plantations radically alters the forest structure and  
347 species composition, modifying ecological processes, and the structure of food webs that can  
348 lead to cascading effects (Brockerhoff *et al.*, 2008; Liebhold *et al.*, 2017). By contrast, native  
349 plantations of locally occurring tree species could serve as substitute habitats for forest beetle  
350 species (Brockerhoff *et al.*, 2008; Méndez-Rojas *et al.*, 2012; Magura *et al.*, 2015). The  
351 similar levels of beetle species richness and/or abundance in natural forest and native forestry  
352 plantations can be explained by the existence of similar habitat structure and plant species  
353 composition, leading to favorable microclimates and food resources (Haddad *et al.*, 2009;  
354 Magura *et al.*, 2015).

355

#### 356 *Effects of the landscape around plantations*

357 Forestry plantations that are small and connected to natural forests show less severe negative  
358 effects on beetle species richness and/or abundance. Local-scale studies have found that  
359 beetle richness or abundance decreased markedly within the surrounding matrix versus  
360 natural forest (Davies & Margules, 1998; Hendrickx *et al.*, 2009; Barnes *et al.*, 2014;  
361 Tóthmérész *et al.*, 2014). Plantations isolated from natural forests may be out of reach for  
362 dispersing beetles, whereas connectivity allows the dispersion of beetles into plantations from  
363 the natural forest (Davis *et al.*, 2001; Gries *et al.*, 2012; Cerda *et al.*, 2015). Higher species  
364 richness in smaller plantations may also be related to the edge effect, enabling the influx of  
365 matrix species into these plantations (González-Vainer *et al.*, 2012; Magura *et al.*, 2017). In  
366 contrast, larger plantations likely exclude both forest specialist and matrix species, reducing  
367 their overall species richness (Magura *et al.*, 2017).

368

### 369 *Effects of plantation age*

370 The success and extent of beetle recolonization fundamentally depend on the origin of the  
371 planted tree species. Favorable environmental conditions, resembling those of the original  
372 forest, may be slower to appear in exotic than native plantations, decreasing the chance of  
373 successful recolonization by native beetle species (Brockerhoff *et al.*, 2008; Nagy *et al.*, 2015).  
374 In exotic plantations, after canopy closure, the number and abundance of open-habitat and  
375 generalist species begins to steeply decline (Pohl *et al.*, 2007; Brockerhoff *et al.*, 2008; Lange  
376 *et al.*, 2014; Nagy *et al.*, 2015). In contrast, canopy closure in native plantations supports the  
377 recruitment of additional native woody plants, makes more intensive litterfall, and thus creates  
378 the microhabitat complexity and favorable microclimates necessary for forest specialist beetle  
379 species (Dent & Wright, 2009; Koivula, 2011; Lange *et al.*, 2014). Thus, successful  
380 recolonization of forest specialist beetle species results in a significant increase in diversity and  
381 abundance in native plantations (Audino *et al.*, 2014; Nagy *et al.*, 2015). Nevertheless, a  
382 complete recovery of native beetle species abundance and diversity is a slow process, taking  
383 >45 years in non-tropical (*e.g.*, Magura *et al.*, 2015) and >60 years in tropical (*e.g.*, Yu *et al.*,  
384 2004, 2006; Sakchoowong *et al.*, 2008; Noriega *et al.*, 2021) biomes.

385

### 386 **Conclusions and conservation recommendations**

387 We found that the replacement of natural forests with forestry plantations negatively affected  
388 the species richness and abundance of three speciose and important beetle families,  
389 Carabidae, Staphylinidae and Scarabaeidae, and that the impact was more severe in forestry  
390 plantations of exotic origin and located in tropical biomes. To retain native forest beetle  
391 communities and associated ecosystem functions and services, particularly in the biodiversity-

392 rich tropical biome, we recommend planting native tree species, allowing them to mature,  
393 maintaining connectivity to natural forests and manage forestry plantations as near-natural  
394 forests. Such silvicultural methods favor mixed species composition and site-adapted tree  
395 species, ensuring the long-term persistence of native forest (Pommerening, 2006; Baral *et al.*,  
396 2016; Pirard *et al.*, 2016).

397         Where continuous vegetation cover is not feasible, we recommend implementing  
398 multi-purpose native plantations rather than using exotic species because beetle abundance  
399 and diversity are better conserved in native plantations (Stephens & Wagner, 2007; Baral *et*  
400 *al.*, 2016). As plantation age significantly increased both species richness and abundance of  
401 beetles in native plantations, we recommend postponing the timber harvest in some stands of  
402 native plantations. We also found that more extensive and more isolated (generally also  
403 exotic) plantations negatively influenced beetle species richness and/or abundance, pointing  
404 to the need for improving landscape-scale connectivity. Thus, **establishing** exotic forestry  
405 plantations in small areas and connected to natural forests could reduce adverse effects on the  
406 diversity of these beetles. **Finally, although in our meta-analysis different countries with**  
407 **forestry tradition located in boreal biomes (e.g., North America, Scandinavian countries, or**  
408 **Russia) were not included because of the lack of necessary statistical data (see results); we**  
409 **conclude that our findings can be applied on these regions based on studies carried out at a**  
410 **local level (e.g., Pohl *et al.*, 2007).**

411

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422

## 423 **References**

424 Audino, L.D., Louzada, J. & Comita, L. (2014). Dung beetles as indicators of tropical forest  
425 restoration success: Is it possible to recover species and functional diversity? *Biological*  
426 *Conservation*, **169**, 248-257.

427 Audino, L.D., Murphy, S.J., Zambaldi, L., Louzada, J. & Comita, L.S. (2017). Drivers of  
428 community assembly in tropical forest restoration sites: role of local environment,  
429 landscape, and space. *Ecological Applications*, **27**, 1731-1745.

430 Baral, H., Guariguata, M.R. & Keenan, R.J. (2016). A proposed framework for assessing  
431 ecosystem goods and services from planted forests. *Ecosystem Services*, **22**, 260-268.

432 Barlow, J., Gardner, T.A., Araujo, I.S., Avila-Pires, T.C., Bonaldo, A.B., Costa, J.E.,  
433 Esposito, M.C., Ferreira, L.V., Hawes, J., Hernandez, M.M., Hoogmoed, M.S., Leite, R.N.,  
434 Lo-Man-Hung, N.F., Malcolm, J.R., Martins, M.B., Mestre, L.A.M., Miranda-Santos, R.,  
435 Nunes-Gutjahr, A.L., Overal, W.L., Parry, L., Peters, S.L., Ribeiro-Junior, M.A., da Silva,  
436 M.N.F., Motta, C.D. & Peres, C.A. (2007a). Quantifying the biodiversity value of tropical

- 437 primary, secondary, and plantation forests. *Proceedings of the National Academy of*  
438 *Sciences of the United States of America*, **104**, 18555-18560.
- 439 Barlow, J., Gardner, T.A., Ferreira, L.V. & Peres, C.A. (2007b). Litter fall and decomposition  
440 in primary, secondary and plantation forests in the Brazilian Amazon. *Forest Ecology and*  
441 *Management*, **247**, 91-97.
- 442 Barnes, A.D., Emberson, R.M., Chapman, H.M., Krell, F.T. & Didham, R.K. (2014). Matrix  
443 habitat restoration alters dung beetle species responses across tropical forest edges.  
444 *Biological Conservation*, **170**, 28-37.
- 445 Begotti, R.A., Pacífico, E.D.S., de Barros-Ferraz, S.F. & Galetti, M. (2018). Landscape  
446 context of plantation forests in the conservation of tropical mammals. *Journal for Nature*  
447 *Conservation*, **41**, 97-105.
- 448 Beiroz, W., Slade, E.M., Barlow, J., Silveira, J.M., Louzada, J. & Sayer, E. (2016). Dung  
449 beetle community dynamics in undisturbed tropical forests: implications for ecological  
450 evaluations of land-use change. *Insect Conservation and Diversity*, **10**, 94-106.
- 451 Borenstein, M., Hedges, L.V., Higgins, J. & Rothstein, H.R. (2005). Comprehensive meta-  
452 analysis Version 2. Biostat, Englewood, NJ.
- 453 Borenstein, M., Hedges, L.V., Higgins, J.P.T. & Rothstein, H.R. (2009). Introduction to meta-  
454 analysis. John Wiley & Sons, Chichester, UK.
- 455 Bowie, M.H., Stokvis, E., Barber, K., Marris, J. & Hodge, S. (2019). Identification of  
456 potential invertebrate bioindicators of restoration trajectory at a quarry site in Hunua,  
457 Auckland, New Zealand. *New Zealand Journal of Ecology*, **43**, 1-11.

- 458 Brockerhoff, E.G., Jactel, H., Parrotta, J.A., Quine, C.P. & Sayer, J. (2008). Plantation forests  
459 and biodiversity: oxymoron or opportunity? *Biodiversity and Conservation*, **17**, 925-951.
- 460 Bull, G.Q., Bazett, M., Schwab, O., Nilsson, S., White, A. & Maginnis, S. (2006). Industrial  
461 forest plantation subsidies: impacts and implications. *Forest Policy Economics*, **9**, 13-31.
- 462 Castaño-Villa, G.J., Estevez, J.V., Guevara, G., Bohada-Murillo, M. & Fontúrbel, F.E.  
463 (2019). Differential effects of forestry plantations on bird diversity: A global assessment.  
464 *Forest Ecology and Management*, **440**, 202-207.
- 465 Cerda, Y., Grez, A.A. & Simonetti, J.A. (2015). The role of the understory on the abundance,  
466 movement and survival of *Ceroglossus chilensis* in pine plantations: an experimental test.  
467 *Journal of Insect Conservation*, **19**, 119-127.
- 468 Chaudhary, A., Burivalova, Z., Koh, L.P. & Hellweg, S. (2016). Impact of forest management  
469 on species richness: Global meta-analysis and economic trade-offs. *Scientific Reports*, **6**,  
470 23954.
- 471 Chazdon, R.L. (2008). Beyond deforestation: restoring forests and ecosystem services on  
472 degraded lands. *Science*, **320**, 1458-1460.
- 473 Chazdon, R.L., Brancalion, P.H., Laestadius, L., Bennett-Curry, A., Buckingham, K., Kumar,  
474 C., Moll-Rocek, J., Vieira, I.M.G. & Wilson, S.J. (2016). When is a forest a forest? Forest  
475 concepts and definitions in the era of forest and landscape restoration. *Ambio*, **45**, 538-550.
- 476 da Silva, P. & Hernández, M.I.M. (2015). Spatial patterns of movement of dung beetle  
477 species in a tropical forest suggest a new trap spacing for dung beetle biodiversity studies.  
478 *PlosOne*, **10**, e0126112.

- 479 Davies, K.F. & Margules, C.R. (1998). Effects of habitat fragmentation on carabid beetles:  
480 experimental evidence. *Journal of Animal Ecology*, **67**, 460-471.
- 481 Davis, A.J., Holloway, J.D., Huijbregts, H., Krikken, J., Kirk-Spriggs, A.H. & Sutton, S.L.  
482 (2001). Dung beetles as indicators of change in the forests of northern Borneo. *Journal of*  
483 *Applied Ecology*, **38**, 593-616.
- 484 Davis, A.L.V., Van Aarde, R.J., Scholtz, C.H. & Delpont, J.H. (2002). Increasing  
485 representation of localized dung beetles across a chronosequence of regenerating  
486 vegetation and natural dune forest in South Africa. *Global Ecology and Biogeography*, **11**,  
487 191-209.
- 488 Dent, D.H. & Wright, J.S. (2009). The future of tropical species in secondary forests: A  
489 quantitative review. *Biological Conservation*, **142**, 2833-2843.
- 490 Derhé, M.A., Murphy, H., Monteith, G. & Menéndez, R. (2016). Measuring the success of  
491 reforestation for restoring biodiversity and ecosystem functioning. *Journal of Applied*  
492 *Ecology*, **53**, 1714-1724.
- 493 Driscoll, D.A., Banks, S.C., Barton, P.S., Lindenmayer, D.B. & Smith, A.L. (2013).  
494 Conceptual domain of the matrix in fragmented landscapes. *Trends in Ecology &*  
495 *Evolution*, **28**, 605-613.
- 496 Duval, S. & Tweedie, R. (2000). Trim and fill: a simple funnel-plot-based method of testing  
497 and adjusting for publication bias in meta-analysis. *Biometrics*, **56**, 455-463.
- 498 Edwards, F.A., Edwards, D.P., Larsen, T.H., Hsu, W.W., Benedick, S., Chung, A., Chey,  
499 V.K., Wilcove, D.S. & Hamer, K.C. (2014). Does logging and forest conversion to oil



- 500 palm agriculture alter functional diversity in a biodiversity hotspot? *Animal Conservation*,  
501 **17**, 163-173.
- 502 Edwards, F.A., Finan, J., Graham, L.K., Larsen, T.H., Wilcove, D.S., Hsu, W.W., Chey, V.K.  
503 & Hamer, K.C. (2017). The impact of logging roads on dung beetle assemblages in a  
504 tropical rainforest reserve. *Biological Conservation*, **205**, 85-92.
- 505 FAO. (2020). Global Forest Resources Assessment 2020 – Key findings. Food and  
506 Agriculture Organization of the United Nations, Rome.
- 507 Fontúrbel, F.E., Candia, A.B., Malebrán, J., Salazar, D.A., González-Browne, C. & Medel, R.  
508 (2015). Meta-analysis of anthropogenic habitat disturbance effects on animal-mediated  
509 seed dispersal. *Global Change Biology*, **21**, 3951-3960.
- 510 Fuzessy, L.F., Benítez-López, A., Slade, E.M., Bufalo, F.S., Magro-de-Souza, G.C., Pereira,  
511 L.A. & Culot, L. (2021). Identifying the anthropogenic drivers of declines in tropical dung  
512 beetle communities and functions. *Biological Conservation*, **256**, 109063
- 513 Gardner, T.A., Hernández, M.I., Barlow, J. & Peres, C.A. (2008). Understanding the  
514 biodiversity consequences of habitat change: the value of secondary and plantation forests  
515 for neotropical dung beetles. *Journal of Applied Ecology*, **45**, 883-893.
- 516 González-Vainer, P., Morelli, E. & Defeo, O. (2012). Differences in coprophilous beetle  
517 communities structure in Sierra de Minas (Uruguay): a mosaic landscape. *Neotropical*  
518 *Entomology*, **41**, 366-374.
- 519 Gries, R., Louzada, J., Almeida, S., Macedo, R. & Barlow, J. (2012). Evaluating the impacts  
520 and conservation value of exotic and native tree afforestation in Cerrado grasslands using  
521 dung beetles. *Insect Conservation and Diversity*, **5**, 175-185.

- 522 Grimbacher, P.S., Catterall, C.P., Kanowski, J. & Proctor, H.C. (2007). Responses of ground-  
523 active beetle assemblages to different styles of reforestation on cleared rainforest land.  
524 *Biodiversity and Conservation*, **16**, 2167-2184.
- 525 Gurevitch, J., Curtis, P.S. & Jones, M.H. (2001). Meta-analysis in ecology. *Advances in*  
526 *Ecological Research*, **32**, 199-247.
- 527 Haddad, N.M., Crutsinger, G.M., Gross, K., Haarstad, J., Knops, J.M.H. & Tilman, D. (2009).  
528 Plant species loss decreases arthropod diversity and shifts trophic structure. *Ecology*  
529 *Letters*, **12**, 1029-1039.
- 530 Halffter, G. (1991). Historical and ecological factors determining the geographical  
531 distribution of beetles (Coleoptera: Scarabaeidae: Scarabaeinae). *Biogeographia—The*  
532 *Journal of Integrative Biogeography*, **15**, 11-40.
- 533 Hanski, I. & Cambefort, Y. (1991). Dung beetle ecology. Princeton University Press, New  
534 Jersey, pp. 520.
- 535 Hedges, L.V. & Olkin, I. (1985). Statistical methods for meta-analysis. Academic Press,  
536 Orlando Florida, pp. 369.
- 537 Hendrickx, F., Maelfait, J.P., Desender, K., Aviron, S., Bailey, D., Diekötter, T., Lens, L.,  
538 Liira, J., Schweiger, O., Speelmans, M., Vandomme, V. & Bugter, R. (2009). Pervasive  
539 effects of dispersal limitation on within-and among-community species richness in  
540 agricultural landscapes. *Global Ecology and Biogeography*, **18**, 607-616.
- 541 Hendrickx, F., Palmer, S.C. & Travis, J.M. (2013). Ideal free distribution of fixed dispersal  
542 phenotypes in a wing dimorphic beetle in heterogeneous landscapes. *Ecology*, **94**, 2487-  
543 2497.

- 544 Henle, K., Davies, K.F., Kleyer, M., Margules, C. & Settele, J. (2004). Predictors of species  
545 sensitivity to fragmentation. *Biodiversity & Conservation*, **13**, 207-251.
- 546 Higgins, J.P.T., Thompson, S.G., Deeks, J.J. & Altman, D.G. (2003). Measuring  
547 inconsistency in meta-analyses. *British Medical Journal*, **327**, 557-560.
- 548 Hillebrand, H. (2008). *Meta-Analysis in Ecology*. John Wiley & Sons, Chichester, UK.
- 549 Horák, J., Brestovanská, T., Mladenović, S., Kout, J., Bogusch, P., Halda, J.P. & Zasadil, P.  
550 (2019). Green desert? Biodiversity patterns in forest plantations. *Forest Ecology and*  
551 *Management*, **433**, 343-348.
- 552 Hozo, S.P., Djulbegovic, B. & Hozo, I. (2005). Estimating the mean and variance from the  
553 median, range, and the size of a sample. *BMC Medical Research Methodology*, **5**, 13-23.
- 554 Humphrey, J.W., Hawes, C., Peace, A.J., Ferris-Kaan, R. & Jukes, M.R. (1999).  
555 Relationships between insect diversity and habitat characteristics in plantation forests.  
556 *Forest Ecology and Management*, **113**, 11-21.
- 557 ITTO. (2016). *Biennial Review and Assessment of the World Timber Situation 2015–2016*.  
558 International Tropical Timber Organization, Yokohama (Japan).
- 559 Jennions, M.D., Lortie, C.J., Rosenberg, M.S. & Rothstein, H.R. (2013). Publication and  
560 related biases. In: Koricheva, J., Gurevitch, J., Mengersen, K. (Eds.), *Handbook of*  
561 *Metaanalysis in Ecology and Evolution*. Princeton University Press, Princeton NJ, pp. 207-  
562 236.
- 563 Kerr, G. (1999). The use of silvicultural systems to enhance the biological diversity of  
564 plantation forests in Britain. *Forestry*, **72**, 191-205.

- 565 Koivula, M. (2011). Useful model organisms, indicators, or both? Ground beetles  
566 (Coleoptera, Carabidae) reflecting environmental conditions. *ZooKeys*, **100**, 287-317.
- 567 Konstantopoulos, S. & Hedges, L.V. (2010). Analyzing effect sizes: Fixed-effects models. In:  
568 H. Cooper, L.V. Hedges y J.C. Valentine (Eds.), *The handbook of research synthesis and*  
569 *meta-analysis 2<sup>a</sup> ed.* pp. 279-293. Nueva York: Russell Sage Foundation.
- 570 Kudavidanage, E.P., Wanger, T.C., De Alwis, C., Sanjeewa, S. & Kotagama, S.W. (2012).  
571 Amphibian and butterfly diversity across a tropical land-use gradient in Sri Lanka;  
572 implications for conservation decision making. *Animal Conservation*, **15**, 253-265.
- 573 Lamb, D., Erskine, P.D. & Parrotta, J.A. (2005). Restoration of degraded tropical forest  
574 landscapes. *Science*, **310**, 1628-1632.
- 575 Lange, M., Türke, M., Pašalić, E., Boch, S., Hessenmöller, D., Müller, J., Prati, D., Socher,  
576 S.A., Fischer, M., Weisser, W.W. & Gossner, M.M. (2014). Effects of forest management  
577 on ground-dwelling beetles (Coleoptera; Carabidae, Staphylinidae) in Central Europe are  
578 mainly mediated by changes in forest structure. *Forest Ecology and Management*, **329**,  
579 166-176.
- 580 Larsen, T.H. & Forsyth, A. (2005). Trap spacing and transect design for dung beetle  
581 biodiversity studies 1. *Biotropica: The Journal of Biology and Conservation*, **37**, 322-325.
- 582 Liebhold, A.M., Brockerhoff, E.G., Kalisz, S., Nuñez, M.A., Wardle, D.A. & Wingfield, M.J.  
583 2017. Biological invasions in forest ecosystems. *Biological Invasions*, **19**, 3437-3458.
- 584 Magura, T., Elek, Z. & Tóthmérész, B. (2002). Impacts of non-native spruce reforestation on  
585 ground beetles. *European Journal of Soil Biology*, **38**, 291-295.

- 586 Magura, T., Bogyó, D., Mizser, S., Nagy, D.D. & Tóthmérész, B. (2015). Recovery of  
587 ground-dwelling assemblages during reforestation with native oak depends on the mobility  
588 and feeding habits of the species. *Forest Ecology and Management*, **339**, 117-126.
- 589 Magura, T., Lövei, G.L. & Tóthmérész, B. (2017). Edge responses are different in edges  
590 under natural versus anthropogenic influence: a meta-analysis using ground beetles.  
591 *Ecology and Evolution*, **7**, 1009-1017.
- 592 Magura, T. & Lövei, L.G. (2019). Environmental filtering is the main assembly rule of  
593 ground beetles in the forest and its edge but not in the adjacent grassland. *Insect Science*,  
594 **26**, 154-163.
- 595 Méndez-Rojas, D.M., López-García, M.M. & Garcia-Cardenas, R. (2012). Diversity of rove  
596 beetles (Coleoptera, Staphylinidae) in restored high-Andean forests from the Colombian  
597 Central Andes. *Revista Colombiana de Entomología*, **38**, 141-147.
- 598 Méndez-Rojas, D.M., Cultid-Medina, C. & Escobar, F. (2021). Influence of land use change  
599 on rove beetle diversity: A systematic review and global meta-analysis of a mega-diverse  
600 insect group. *Ecological Indicators*, **122**, 107239.
- 601 Mengersen, K., Jennions, M.D. & Schmid, C. (2013). Statistical models for the meta-analysis  
602 of non-independent data. In: Koricheva, J., Gurevitch, J., Mengersen, K. (Eds.), Handbook  
603 of meta-analysis in ecology and evolution. Princeton University Press, New Jersey, pp.  
604 255-264.
- 605 Milheiras, S.G., Guedes, M., Silva, F.A.B., Aparício, P. & Mace, G.M. (2020). Patterns of  
606 biodiversity response along a gradient of forest use in Eastern Amazonia, Brazil. *PeerJ*, **8**,  
607 e8486.

- 608 Moher, D., Liberati, A., Tetzlaff, J., Altman, D.G. & Prisma Group. (2009). Preferred  
609 reporting items for systematic reviews and meta-analyses: the PRISMA statement. *PLoS*  
610 *Medicine*, **6**, e1000097.
- 611 Nagy, D.D., Magura, T., Debnár, Z., Horváth, R. & Tóthmérész, B. (2015). Shift of rove  
612 beetle assemblages in reforestations: Does nativity matter? *Journal of Insect Conservation*,  
613 **19**, 1075-1087.
- 614 Nakagawa, S., Noble, D.W., Senior, A.M. & Lagisz, M. (2017). Meta-evaluation of meta-  
615 analysis: ten appraisal questions for biologists. *Biomedical Central Biology*, **15**, 1-14.
- 616 New, T.R. (2010). Beetles in conservation. Wiley-Blackwell, Oxford, UK.
- 617 Newbold, T., Hudson, L.N., Hill, S.L.L., Contu, S., Lysenko, I., Senior, R.A., Börger, L.,  
618 Bennett, D.J., Choimes, A., Collen, B., Day, J., De Palma, A., Díaz, A., Echeverria-  
619 Londoño, S., Edgar, M.J., Feldman, A., Garon, M., Harrison, M.L.K., Alhusseini, T.,  
620 Ingram, D.J., Itescu, Y., Kattge, J., Kemp, V., Kirkpatrick, L., Kleyer, M., Correia, D.L.P.,  
621 Martin, C.D., Meiri, S., Novosolov, M., Pan, Y., Phillips, H.R.P., Purves, D.W., Robinson,  
622 A., Simpson, J., Tuck, S.L., Weiher, E., White, H.J., Ewers, R.M., Mace, G.M.,  
623 Scharlemann, J.P.W. & Purvis, A. (2015). Global effects of land use on local terrestrial  
624 biodiversity. *Nature*, **520**, 45-50.
- 625 Nichols, E., Larsen, T., Spector, S., Davis, A.L., Escobar, F., Favila, M., Vulinec, K. &  
626 Network, T.S.R. (2007). Global dung beetle response to tropical forest modification and  
627 fragmentation: a quantitative literature review and meta-analysis. *Biological Conservation*,  
628 **137**, 1-19.

- 629 Nichols, E., Spector, S., Louzada, J., Larsen, T., Amezquita, S., Favila, M.E., & Network,  
630 T.S.R. (2008). Ecological functions and ecosystem services provided by Scarabaeinae  
631 dung beetles. *Biological Conservation*, **141**, 1461-1474.
- 632 Niemelä, J., Koivula, M.D. & Kotze, J. (2007). The effects of forestry on carabid beetles  
633 (Coleoptera: Carabidae) in boreal forests. *Journal of Insect Conservation*, **11**, 5-18.
- 634 Noriega, J.A., Santos, A.M.C., Calatayud, J., Chozas, S. & Hortal, J. (2021). Short- and long-  
635 term temporal changes in the assemblage structure of Amazonian dung beetles. *Oecologia*,  
636 **195**, 719-736.
- 637 Paillet, Y., Bergès, L., Hjältén, J., Odor, P., Avon, C., Bernhardt-Römermann, M., Bijlsma,  
638 R.J, De Bruyn, L., Fuhr, M., Grandin, U., Kanka, R., Lundin, L., Luque, S., Magura, T.,  
639 Matesanz, S., Mészáros, I., Sebastià, M.T., Schmidt, W., Standovár, T., Tóthmérész, B.,  
640 Uotila, A., Valladares, F., Vellak, K. & Virtanen, R. (2010). Biodiversity differences  
641 between managed and unmanaged forests: meta-analysis of species richness in Europe.  
642 *Conservation Biology*, **24**, 101-112.
- 643 Paquette, A. & Messier, C. (2010). The role of plantations in managing the world's forests in  
644 the Anthropocene. *Frontiers in Ecology and the Environment*, **8**, 27-34.
- 645 Paul, K.L., Jacobsen, K., Koul, V., Leppert, P. & Smith, J. (2008). Predicting growth and  
646 sequestration of carbon by plantations growing in regions of low-rainfall in southern  
647 Australia. *Forest Ecology and Management*, **254**, 205-216.
- 648 Payn, T., Carnus, J.M., Freer-Smith, P., Kimberley, M., Kollert, W., Liu, S., Orazio, C.,  
649 Rodriguez, L., Neves, L. & Wingfield, M.J. (2015). Changes in planted forests and future  
650 global implications. *Forest Ecology and Management*, **352**, 57-67.

- 651 Piotto, D. (2008). A meta-analysis comparing tree growth in monocultures and mixed  
652 plantations. *Forest Ecology and Management*, **255**, 781-786.
- 653 Pirard, R., Dal Secco, L. & Warman, R. (2016). Do timber plantations contribute to forest  
654 conservation? *Environmental Science & Policy*, **57**, 122-130.
- 655 [Plexida, S.G., Sfougaris, A.I., & Papadopoulos, N.T. \(2014\). The impact of human land use  
656 on the composition and richness of ground and dung beetle assemblages. \*Applied Ecology  
657 and Environmental Research\*, \*\*12\*\*, 661-679.](#)
- 658 Pohl, G.R., Langor, D.W. & Spence, J.R. (2007). Rove beetles and ground beetles  
659 (Coleoptera: Staphylinidae, Carabidae) as indicators of harvest and regeneration practices  
660 in western Canadian foothills forests. *Biological Conservation*, **137**, 294-307.
- 661 Pohl, G.R., Langor, D.W., Klimaszewski, J., Work, T. & Paquin, P. (2008). Rove beetles  
662 (Coleoptera: Staphylinidae) in northern Nearctic forests. *The Canadian Entomologist*, **140**,  
663 415-436.
- 664 Pommerening, A. (2006). Transformation to continuous cover forestry in a changing  
665 environment. *Forest Ecology and Management*, **224**, 227-228.
- 666 Proença, V.M., Pereira, H.M., Guilherme, J. & Vicente, L. (2010). Plant and bird diversity in  
667 natural forests and in native and exotic plantations in NW Portugal. *Acta Oecologica*, **36**,  
668 219-226.
- 669 Rainio, J. & Niemelä, J. (2003). Ground beetles (Coleoptera: Carabidae) as bioindicators.  
670 *Biodiversity & Conservation*, **12**, 487-506.



- 671 Roberge, J.M. & Stenbacka, F. (2014). Assemblages of epigeaic beetles and understory  
672 vegetation differ between stands of an introduced pine and its native congener in boreal  
673 forest. *Forest Ecology and Management*, **318**, 239-249.
- 674 Robson, T.C., Baker, A.C. & Murray, B.R. (2009). Differences in leaf-litter invertebrate  
675 assemblages between radiata pine plantations and neighbouring native eucalypt woodland.  
676 *Austral Ecology*, **34**, 368-376.
- 677 Sakchoowong, W., Nomura, S., Ogata, K. & Chanpaisaeng, J. (2008). Diversity of pselaphine  
678 beetles (Coleoptera: Staphylinidae: Pselaphinae) in eastern Thailand. *Entomological*  
679 *Science*, **11**, 301-313.
- 680 Samways, M.J. (2005). Insect diversity conservation. Cambridge University Press,  
681 Cambridge, UK.
- 682 Schowalter, T.D. (2008). Insects and sustainability of ecosystem services. CRC Press, Taylor  
683 & Francis group, New York, USA.
- 684 Senior, R.A., Hill, J.K., González del Pliego, P., Goode, L.K. & Edwards, D.P. (2017). A  
685 pantropical analysis of the impacts of forest degradation and conversion on local  
686 temperature. *Ecology and Evolution*, **7**, 7897-7908.
- 687 Simonetti, J.A., Grez, A.A. & Estades, C.F. (2012). Biodiversity conservation in agroforestry  
688 landscapes: Challenges and opportunities. Editorial Universitaria, Santiago de Chile.
- 689 Spector, S. (2006). Scarabaeinae dung beetles (Coleoptera: Scarabaeidae: Scarabaeinae): An  
690 invertebrate focal taxon for biodiversity research and conservation. *The Coleopterists*  
691 *Bulletin*, **60**, 71-83.

- 692 Stephens, S.S. & Wagner, M.R. (2007). Forest plantations and biodiversity: a fresh  
693 perspective. *Journal of Forestry*, **105**, 307-313.
- 694 Taboada, Á., Tárrega, R., Calvo, L., Marcos, E., Marcos, J.A. & Salgado, J.M. (2010). Plant  
695 and carabid beetle species diversity in relation to forest type and structural heterogeneity.  
696 *European Journal of Forest Research*, **129**, 31-45.
- 697 Thompson, S.G. & Higgins, J.P. (2002). How should meta-regression analyses be undertaken  
698 and interpreted? *Statistics in Medicine*, **21**, 1559-1573.
- 699 Tóthmérész, B., Nagy, D.D., Mizser, S., Bogyó, D. & Magura, T. (2014). Edge effects on  
700 ground-dwelling beetles (Carabidae and Staphylinidae) in oak forest-forest edge-grassland  
701 habitats in Hungary. *European Journal of Entomology*, **111**, 686-691.
- 702 Wan, X., Wang, W., Liu, J. & Tong, T. (2014). Estimating the sample mean and standard  
703 deviation from the sample size, median, range and/or interquartile range. *BMC Medical  
704 Research Methodology*, **14**, 135-149.
- 705 Yu, X.D., Luo, T.H. & Zhou, H.Z. (2004). *Carabus* (Coleoptera: Carabidae) assemblages of  
706 native forests and non-native plantations in Northern China. *Entomologica Fennica*, **15**,  
707 129-137.
- 708 Yu, X.D., Luo, T.H. & Zhou, H.Z. (2006). Distribution of carabid beetles among regenerating  
709 and natural forest types in Southwestern China. *Forest Ecology and Management*, **231**,  
710 169-177.
- 711

## 712 **Figure legends**

713 **Fig. 1** Location of the 48 primary studies, by country, that were used in the meta-analysis.

714 Some sampling sites include more than one article performed in different years.

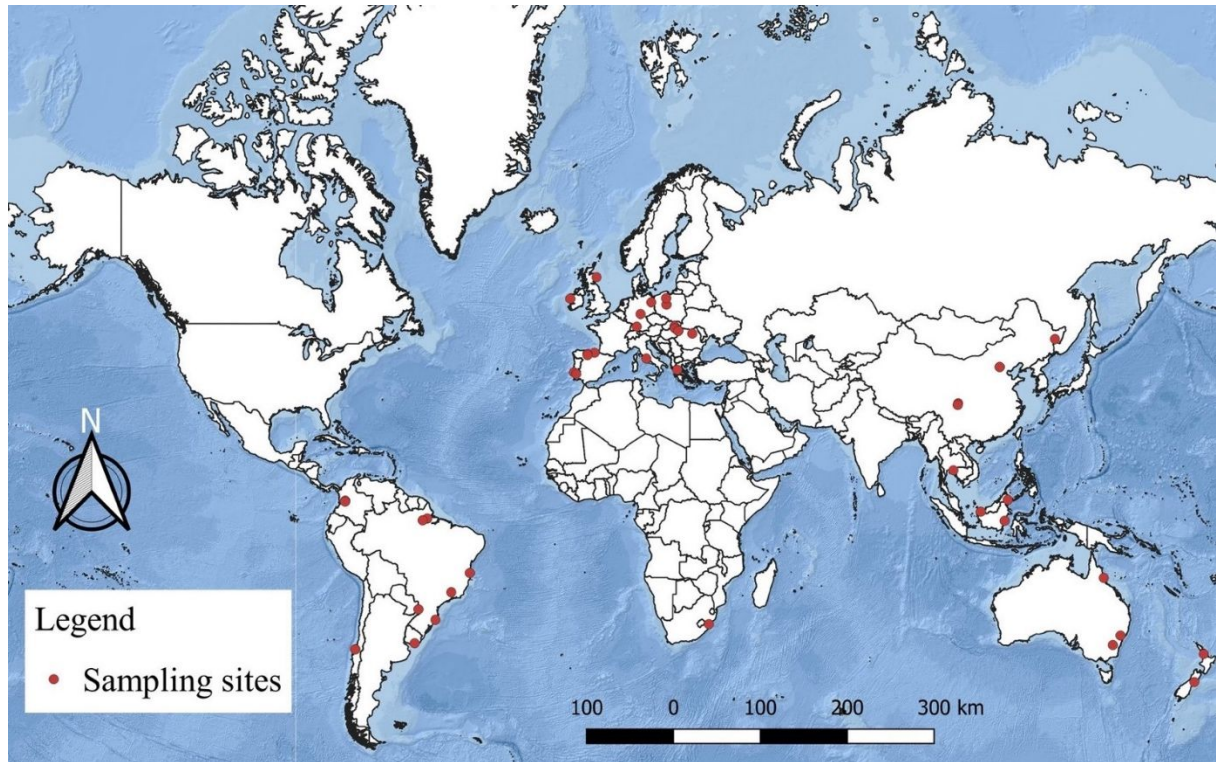
715 **Fig. 2** Effect of natural forest conversion to forestry plantations on beetle species richness.

716 Average values and 95% confidence intervals are given for: (a) the overall effect, (b)  
717 taxonomic group, (c) origin of the planted species, (d) biome type, (e) interaction between  
718 species origin and biome type, (f) genus of the planted species, (g) plantation composition, (h)  
719 plantation's purpose, (i) plantation connectivity, and (j) plantation size. The number of case  
720 studies for each level of the moderator variable is in parenthesis. Asterisks indicate that  
721 confidence intervals are significantly different from zero ( $^{NS}p \geq 0.05$ ,  $*p < 0.05$ ,  $**p < 0.01$ ,  
722  $***p < 0.001$ ).  $Q_b$  represents the homogeneity in group comparisons.

723 **Fig. 3** Effect of natural forest conversion to forestry plantations on beetle abundance. Average  
724 values and 95% confidence intervals are given for: (a) the overall effect, (b) taxonomic group,  
725 (c) origin of the planted species, (d) biome type, (e) interaction between species origin and  
726 biome type, (f) genus of the planted species, (g) plantation composition, (h) plantation's  
727 purpose, (i) plantation connectivity, and (j) plantation size. The number of case studies for  
728 each level of the moderator variable is in parenthesis. Asterisks indicate that confidence  
729 intervals are significantly different from zero ( $^{NS}p \geq 0.05$ ,  $*p < 0.05$ ,  $**p < 0.01$ ,  $***p <$   
730  $0.001$ ).  $Q_b$  represents the homogeneity in group comparisons.

731 **Fig. 4** Relationship between plantation age and beetle species richness (a, b) and abundance  
732 (c, d) in plantations with native (a, c) and exotic (b, d) tree species.

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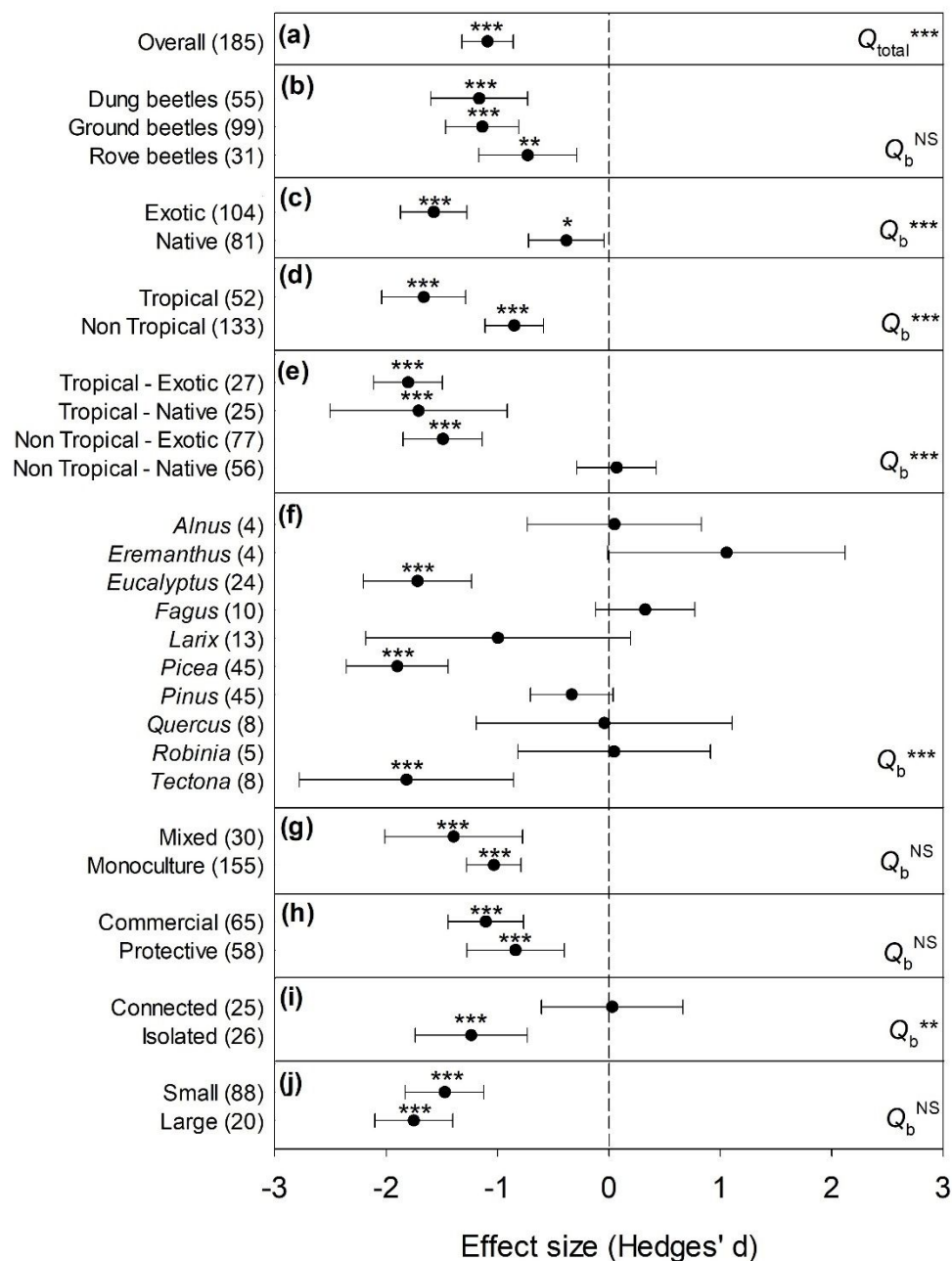
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735

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743 species origin and biome type, (f) genus of the planted species, (g) plantation composition, (h)

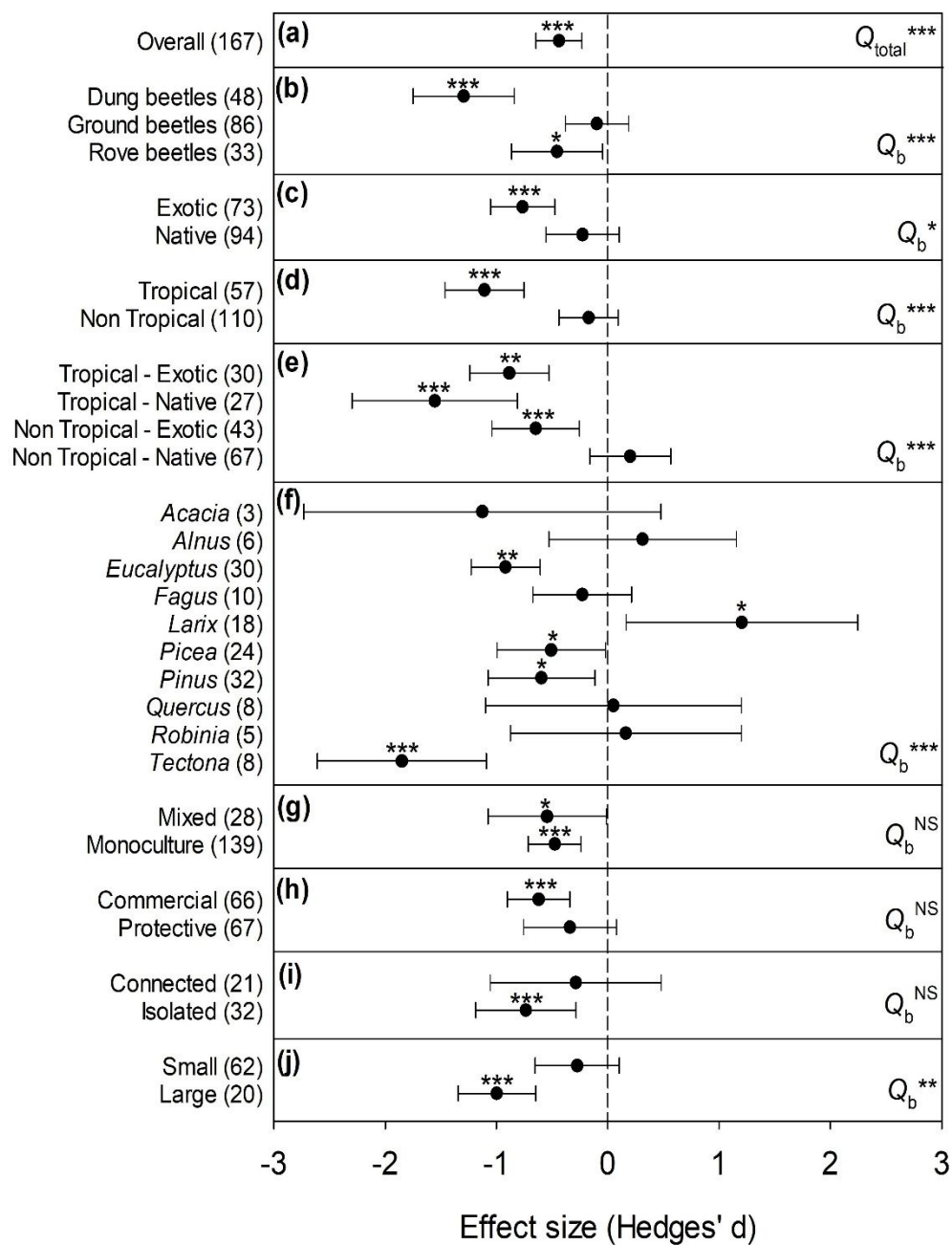
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747  $***p < 0.001$ ).  $Q_b$  represents the homogeneity in group comparisons.

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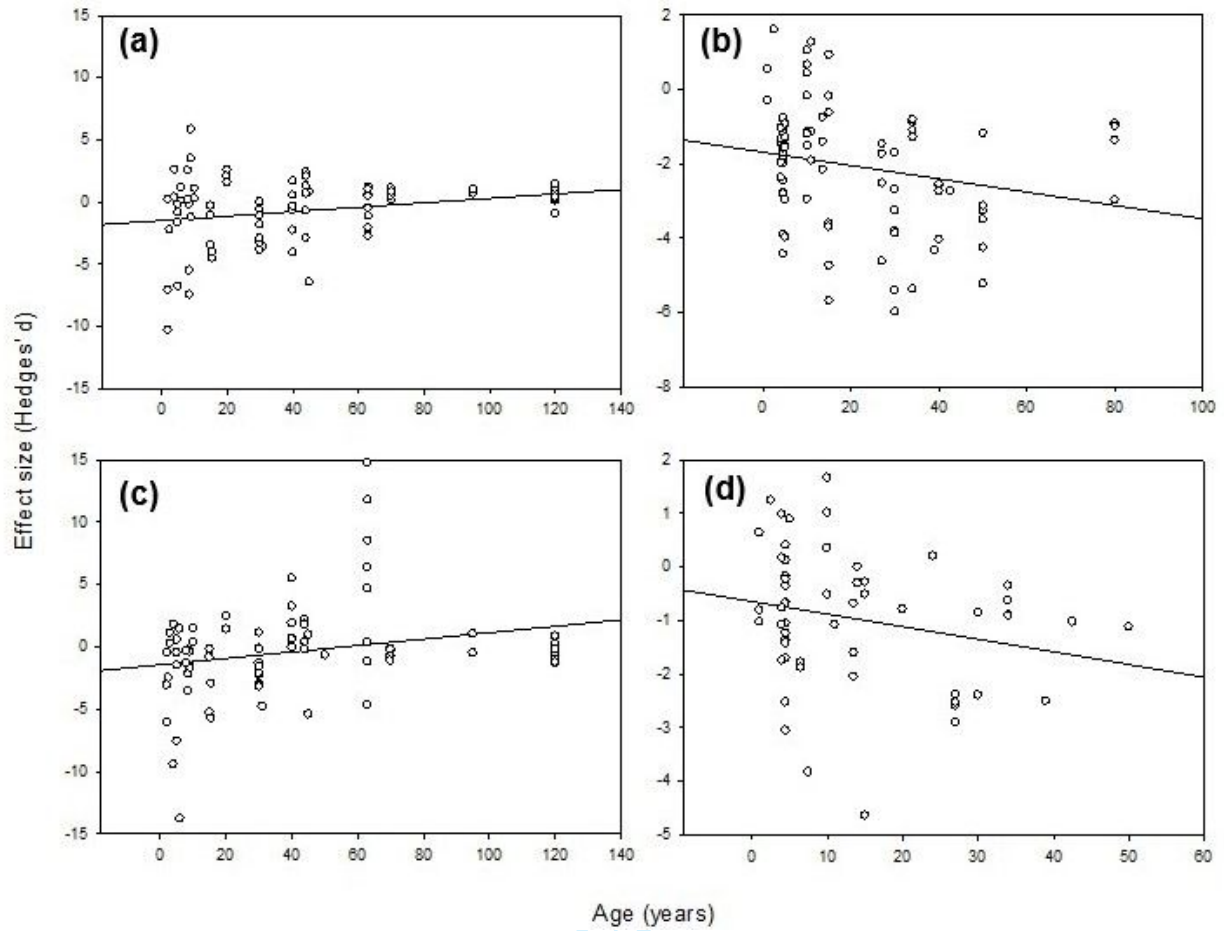
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 752 (c) origin of the planted species, (d) biome type, (e) interaction between species origin and  
 753 biome type, (f) genus of the planted species, (g) plantation composition, (h) plantation's  
 754 purpose, (i) plantation connectivity, and (j) plantation size. The number of case studies for

755 each level of the moderator variable is in parenthesis. Asterisks indicate that confidence  
756 intervals are significantly different from zero (<sup>NS</sup> $p \geq 0.05$ , \* $p < 0.05$ , \*\* $p < 0.01$ , \*\*\* $p <$   
757  $0.001$ ).  $Q_b$  represents the homogeneity in group comparisons.

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760 **Fig. 4** Relationship between plantation age and beetle species richness (a, b) and abundance

761 (c, d) in plantations with native (a, c) and exotic (b, d) tree species.

## SUPPLEMENTARY MATERIAL

**What level of native beetle diversity can be supported by forestry plantations? A global synthesis**

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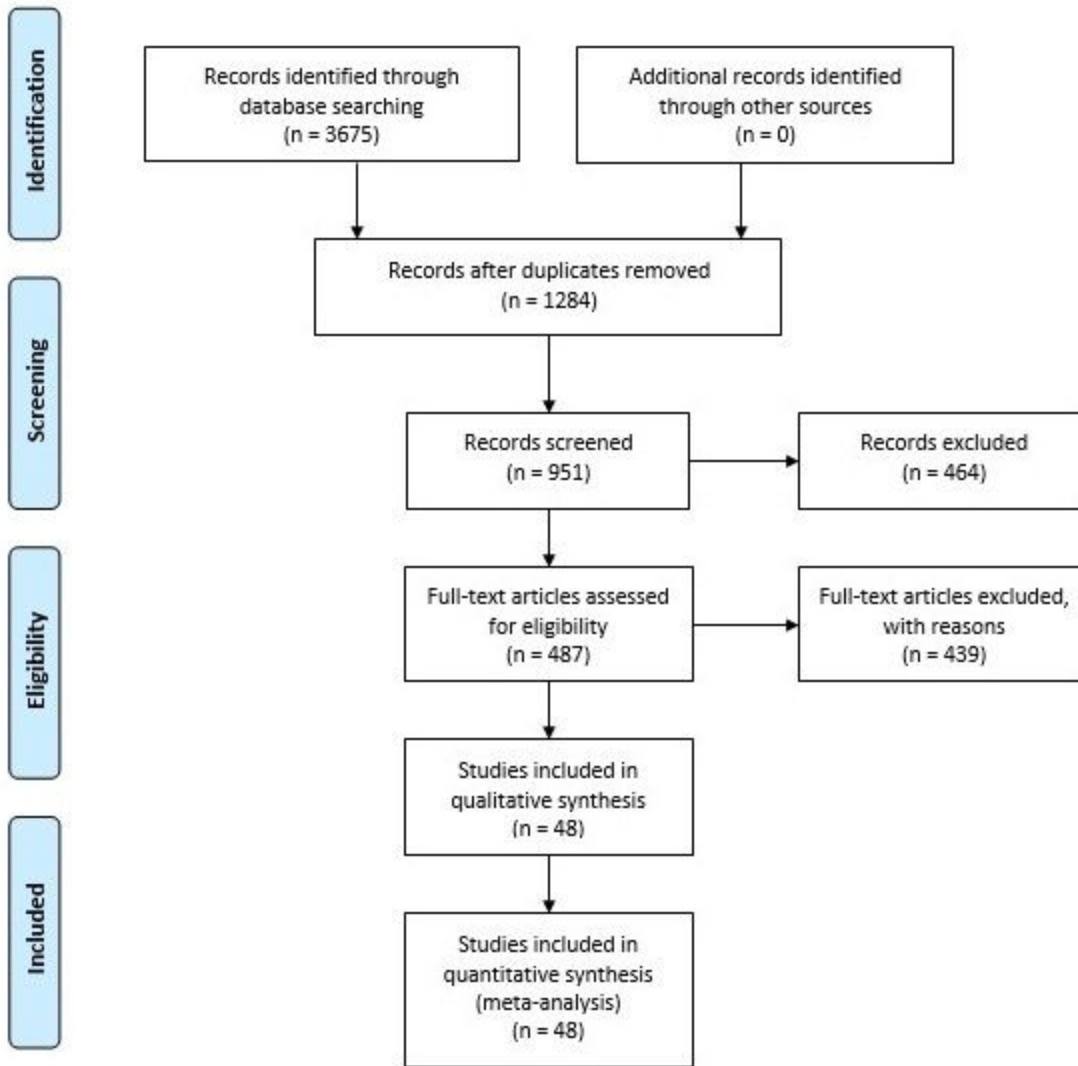
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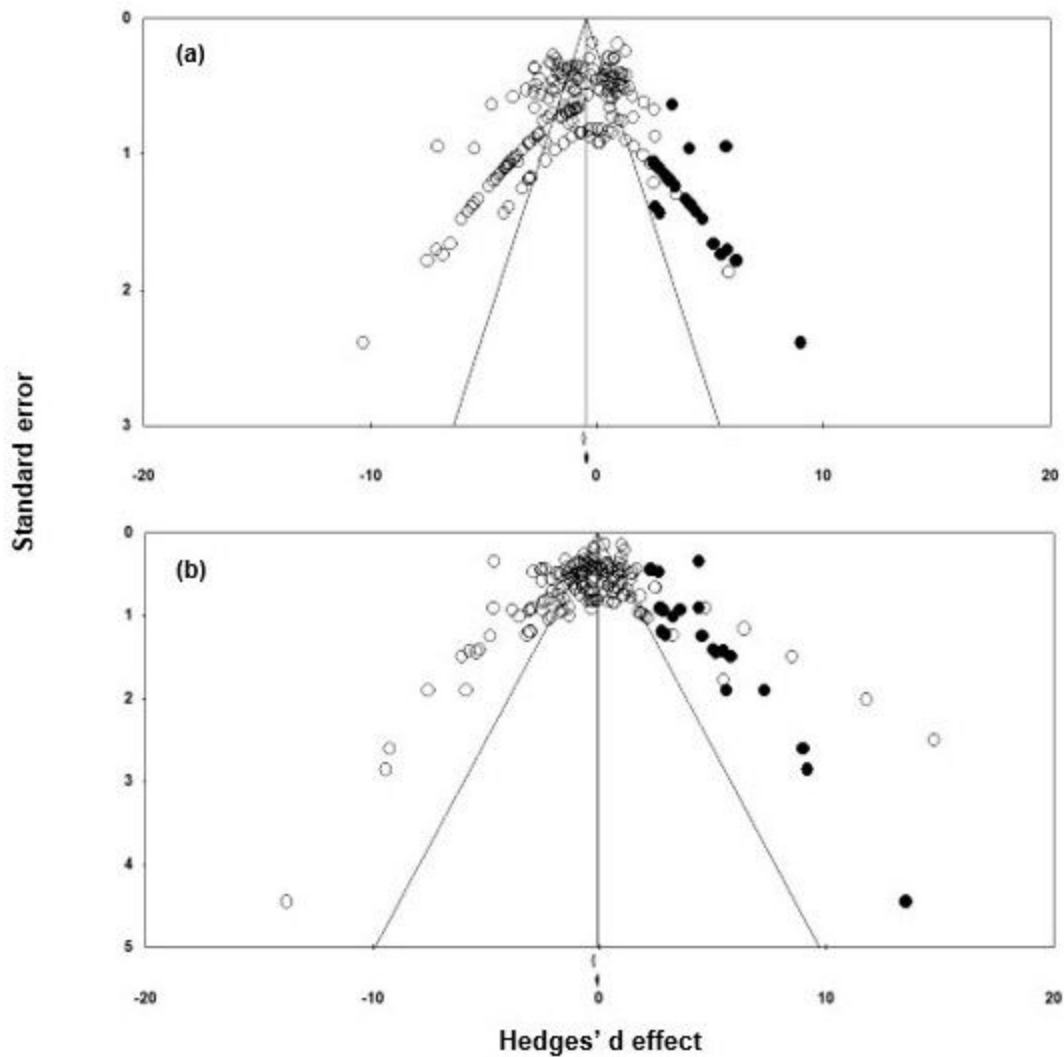
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**Fig. S1.** Flow diagram according to the PRISMA methodology (Preferred Reporting Items for Systematic Reviews and Meta-analysis) carried out in this study.



**Fig. S2.** Funnel plot for beetle species richness (a), and abundance (b). Open circles represent observed case studies and solid circles represent imputed case studies after trim and fill procedures. Open diamonds indicate the observed effect size (mean point estimate), and solid diamonds indicate the adjusted effect size.

**Table S1.** Summary information of primary studies used in the meta-analysis, detailing the assessed metrics, beetle group, origin of planted species, biome type, plantation composition, plantation purpose, plantation connectivity, plantation size, plantation age (years), and planted species. Some primary studies provided more than one case study, namely different sampling locations or different sampling time, which were considered as independent cases. Hedge's  $d$  and variance values (Var) are reported.

Ref.	Metric	Group	Origin	Biome	Composition	Purpose	Landscape	Age	Size	Tree genera	Hedge's $d$	Var
1	Richness	Dung beetle	Native	Tropical	Mixed	Protective	No data	15,5	< 400 ha	No data	-4,536	1,429
1	Richness	Dung beetle	Native	Tropical	Mixed	Protective	No data	8,5	< 400 ha	No data	-7,467	3,188
1	Richness	Dung beetle	Native	Tropical	Mixed	Protective	No data	2	< 400 ha	No data	-10,309	5,713
1	Richness	Dung beetle	Native	Tropical	Mixed	Protective	No data	15,5	< 400 ha	No data	-4,026	1,210
1	Richness	Dung beetle	Native	Tropical	Mixed	Protective	No data	8,5	< 400 ha	No data	-5,515	1,921
1	Richness	Dung beetle	Native	Tropical	Mixed	Protective	No data	2	< 400 ha	No data	-7,071	2,900
1	Abundance	Dung beetle	Native	Tropical	Mixed	Protective	No data	15,5	< 400 ha	No data	-5,740	2,047
1	Abundance	Dung beetle	Native	Tropical	Mixed	Protective	No data	8,5	< 400 ha	No data	-3,533	1,024
1	Abundance	Dung beetle	Native	Tropical	Mixed	Protective	No data	2	< 400 ha	No data	-6,053	2,232
1	Abundance	Dung beetle	Native	Tropical	Mixed	Protective	No data	15,5	< 400 ha	No data	-2,950	0,835
1	Abundance	Dung beetle	Native	Tropical	Mixed	Protective	No data	8,5	< 400 ha	No data	-2,199	0,642
1	Abundance	Dung beetle	Native	Tropical	Mixed	Protective	No data	2	< 400 ha	No data	-3,086	0,876
2	Richness	Ground beetle	Native	Non-tropical	Monoculture	Protective	Connected	20	< 400 ha	<i>Quercus</i>	2,537	0,451
2	Richness	Ground beetle	Native	Non-tropical	Monoculture	Protective	Connected	20	< 400 ha	<i>Quercus</i>	1,560	0,326
2	Richness	Ground beetle	Native	Non-tropical	Monoculture	Protective	Connected	20	< 400 ha	<i>Quercus</i>	2,063	0,383
2	Abundance	Ground beetle	Native	Non-tropical	Monoculture	Protective	Connected	20	< 400 ha	<i>Quercus</i>	2,452	0,438
2	Abundance	Ground beetle	Native	Non-tropical	Monoculture	Protective	Connected	20	< 400 ha	<i>Quercus</i>	1,410	0,312
2	Abundance	Ground beetle	Native	Non-tropical	Monoculture	Protective	Connected	20	< 400 ha	<i>Quercus</i>	2,480	0,442
3	Richness	Ground beetle	Native	Non-tropical	Mixed	No data	No data	No data	< 400 ha	<i>Pinus</i>	0,045	0,200
3	Abundance	Ground beetle	Native	Non-tropical	Mixed	No data	No data	No data	< 400 ha	<i>Pinus</i>	0,090	0,200
4	Richness	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Isolated	4,5	> 1000 ha	<i>Eucalyptus</i>	-1,759	0,259
4	Richness	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Isolated	4,5	> 1000 ha	<i>Eucalyptus</i>	-2,781	0,293
4	Richness	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Isolated	4,5	> 1000 ha	<i>Eucalyptus</i>	-2,448	0,322
4	Richness	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Isolated	4,5	> 1000 ha	<i>Eucalyptus</i>	-1,891	0,216
4	Richness	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Isolated	4,5	> 1000 ha	<i>Eucalyptus</i>	-1,979	0,277
4	Richness	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Isolated	4,5	> 1000 ha	<i>Eucalyptus</i>	-2,805	0,296
4	Richness	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Isolated	4,5	> 1000 ha	<i>Eucalyptus</i>	-1,149	0,220
4	Richness	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Isolated	4,5	> 1000 ha	<i>Eucalyptus</i>	-1,014	0,169
4	Richness	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Isolated	4,5	> 1000 ha	<i>Eucalyptus</i>	-0,818	0,206
4	Richness	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Isolated	4,5	> 1000 ha	<i>Eucalyptus</i>	-1,468	0,190
4	Abundance	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Isolated	4,5	> 1000 ha	<i>Eucalyptus</i>	-1,368	0,232
4	Abundance	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Isolated	4,5	> 1000 ha	<i>Eucalyptus</i>	-1,239	0,178

4	Abundance	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Isolated	4,5	> 1000 ha	<i>Eucalyptus</i>	-1,710	0,255
4	Abundance	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Isolated	4,5	> 1000 ha	<i>Eucalyptus</i>	-0,673	0,158
4	Abundance	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Isolated	4,5	> 1000 ha	<i>Eucalyptus</i>	-2,530	0,331
4	Abundance	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Isolated	4,5	> 1000 ha	<i>Eucalyptus</i>	-1,424	0,188
4	Abundance	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Isolated	4,5	> 1000 ha	<i>Eucalyptus</i>	-0,177	0,192
4	Abundance	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Isolated	4,5	> 1000 ha	<i>Eucalyptus</i>	-0,361	0,152
4	Abundance	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Isolated	4,5	> 1000 ha	<i>Eucalyptus</i>	0,123	0,192
4	Abundance	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Isolated	4,5	> 1000 ha	<i>Eucalyptus</i>	-0,224	0,151
5	Abundance	Ground beetle	Exotic	Non-tropical	Monoculture	Commercial	Isolated	4	No data	<i>Pinus</i>	0,985	0,022
5	Abundance	Ground beetle	Exotic	Non-tropical	Monoculture	Commercial	Isolated	24	No data	<i>Pinus</i>	0,201	0,020
6	Abundance	Rove beetle	Native	Non-tropical	Mixed	Protective	Connected	3	< 400 ha	No data	1,093	0,575
6	Abundance	Ground beetle	Native	Non-tropical	Mixed	Protective	Connected	3	< 400 ha	No data	0,259	0,504
7	Richness	Ground beetle	Native	Non-tropical	Monoculture	Protective	No data	63	No data	<i>Larix</i>	-0,503	0,229
7	Richness	Ground beetle	Native	Non-tropical	Monoculture	Protective	No data	63	No data	<i>Pinus</i>	-1,154	0,259
7	Richness	Ground beetle	Native	Non-tropical	Monoculture	Protective	No data	63	No data	<i>Larix</i>	1,209	0,263
7	Richness	Ground beetle	Native	Non-tropical	Monoculture	Protective	No data	63	No data	<i>Pinus</i>	0,598	0,232
7	Richness	Ground beetle	Native	Non-tropical	Monoculture	Protective	No data	63	No data	<i>Larix</i>	0,962	0,248
7	Richness	Ground beetle	Native	Non-tropical	Monoculture	Protective	No data	63	No data	<i>Pinus</i>	0,472	0,228
7	Richness	Ground beetle	Native	Non-tropical	Monoculture	Protective	No data	63	No data	<i>Larix</i>	-2,127	0,348
7	Richness	Ground beetle	Native	Non-tropical	Monoculture	Protective	No data	63	No data	<i>Pinus</i>	-2,758	0,434
7	Abundance	Ground beetle	Native	Non-tropical	Monoculture	Protective	No data	63	No data	<i>Larix</i>	6,368	1,349
7	Abundance	Ground beetle	Native	Non-tropical	Monoculture	Protective	No data	63	No data	<i>Pinus</i>	-4,665	0,827
7	Abundance	Ground beetle	Native	Non-tropical	Monoculture	Protective	No data	63	No data	<i>Larix</i>	14,748	6,264
7	Abundance	Ground beetle	Native	Non-tropical	Monoculture	Protective	No data	63	No data	<i>Pinus</i>	4,661	0,826
7	Abundance	Ground beetle	Native	Non-tropical	Monoculture	Protective	No data	63	No data	<i>Larix</i>	11,760	4,064
7	Abundance	Ground beetle	Native	Non-tropical	Monoculture	Protective	No data	63	No data	<i>Pinus</i>	0,370	0,226
7	Abundance	Ground beetle	Native	Non-tropical	Monoculture	Protective	No data	63	No data	<i>Larix</i>	8,503	2,231
7	Abundance	Ground beetle	Native	Non-tropical	Monoculture	Protective	No data	63	No data	<i>Pinus</i>	-1,207	0,263
8	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	Commercial	No data	No data	No data	<i>Eucalyptus</i>	-7,010	0,893
8	Abundance	Ground beetle	Exotic	Non-tropical	Monoculture	Commercial	No data	No data	No data	<i>Eucalyptus</i>	-0,509	0,129
9	Richness	Dung beetle	Exotic	Tropical	Mixed	Commercial	Connected	No data	No data	<i>Sweitenia and Acacia</i>	-2,745	0,130
9	Richness	Dung beetle	Exotic	Tropical	Mixed	Commercial	Connected	No data	No data	<i>Sweitenia and Acacia</i>	-1,774	0,091
9	Richness	Dung beetle	Exotic	Tropical	Mixed	Commercial	Connected	No data	No data	<i>Sweitenia and Acacia</i>	-1,927	0,113
9	Richness	Dung beetle	Exotic	Tropical	Mixed	Commercial	Connected	No data	No data	<i>Sweitenia and Acacia</i>	-2,083	0,108
9	Richness	Dung beetle	Exotic	Tropical	Mixed	Commercial	Connected	No data	No data	<i>Sweitenia and Acacia</i>	-2,736	0,136
10	Richness	Dung beetle	Native	Tropical	Mixed	Protective	No data	2,5	< 400 ha	No data	-2,273	0,823
10	Richness	Dung beetle	Native	Tropical	Mixed	Protective	No data	9	< 400 ha	No data	-1,247	0,597
10	Richness	Dung beetle	Native	Tropical	Mixed	Protective	No data	15	< 400 ha	No data	-1,070	0,572
10	Abundance	Dung beetle	Native	Tropical	Mixed	Protective	No data	2,5	< 400 ha	No data	-2,489	0,887
10	Abundance	Dung beetle	Native	Tropical	Mixed	Protective	No data	9	< 400 ha	No data	-1,549	0,650
10	Abundance	Dung beetle	Native	Tropical	Mixed	Protective	No data	15	< 400 ha	No data	-0,855	0,546

Insect Conservation and Diversity

11	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	Protective	No data	34	No data	<i>Picea</i>	-5,380	0,924
12	Richness	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Isolated	4,5	> 1000 ha	<i>Eucalyptus</i>	-4,441	1,386
12	Richness	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Isolated	4,5	> 1000 ha	<i>Eucalyptus</i>	-1,593	0,527
12	Richness	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Isolated	4,5	> 1000 ha	<i>Eucalyptus</i>	-3,910	1,164
12	Richness	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Isolated	4,5	> 1000 ha	<i>Eucalyptus</i>	-0,769	0,430
12	Abundance	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Isolated	4,5	> 1000 ha	<i>Eucalyptus</i>	-1,060	0,456
12	Abundance	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Isolated	4,5	> 1000 ha	<i>Eucalyptus</i>	0,408	0,408
12	Abundance	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Isolated	4,5	> 1000 ha	<i>Eucalyptus</i>	-3,057	0,867
12	Abundance	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Isolated	4,5	> 1000 ha	<i>Eucalyptus</i>	-0,676	0,423
13	Richness	Dung beetle	Exotic	Non-tropical	Monoculture	Commercial	No data	11	No data	<i>Pinus</i>	1,271	0,060
14	Richness	Dung beetle	Exotic	Non-tropical	Monoculture	Commercial	No data	11	No data	<i>Pinus</i>	-1,919	0,073
15	Richness	Dung beetle	Native	Non-tropical	Mixed	Protective	No data	2	No data	<i>Eucalyptus</i>	0,222	0,671
15	Richness	Dung beetle	Native	Non-tropical	Mixed	Protective	No data	8,5	No data	<i>Eucalyptus</i>	-0,219	0,671
15	Abundance	Dung beetle	Native	Non-tropical	Mixed	Protective	No data	2	No data	<i>Eucalyptus</i>	-0,509	0,688
15	Abundance	Dung beetle	Native	Non-tropical	Mixed	Protective	No data	8,5	No data	<i>Eucalyptus</i>	-0,300	0,674
16	Richness	Dung beetle	Exotic	Non-tropical	Monoculture	No data	Isolated	15	< 400 ha	<i>Pinus</i>	0,924	0,036
16	Richness	Rove beetle	Exotic	Non-tropical	Monoculture	No data	Isolated	15	< 400 ha	<i>Pinus</i>	-0,180	0,032
16	Abundance	Dung beetle	Exotic	Non-tropical	Monoculture	No data	Isolated	15	< 400 ha	<i>Pinus</i>	-4,645	0,119
16	Abundance	Rove beetle	Exotic	Non-tropical	Monoculture	No data	Isolated	15	< 400 ha	<i>Pinus</i>	-0,283	0,033
17	Richness	Dung beetle	Native	Tropical	Monoculture	Commercial	Connected	4	No data	<i>Eremanthus</i>	0,364	0,761
17	Richness	Dung beetle	Native	Tropical	Monoculture	Commercial	Connected	6	No data	<i>Eremanthus</i>	0,063	0,834
17	Richness	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Connected	1	No data	<i>Eucalyptus</i>	-0,290	0,757
17	Richness	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Connected	4	No data	<i>Eucalyptus</i>	-1,443	0,849
17	Richness	Dung beetle	Native	Tropical	Monoculture	Commercial	Connected	4	No data	<i>Eremanthus</i>	2,582	0,750
17	Richness	Dung beetle	Native	Tropical	Monoculture	Commercial	Connected	6	No data	<i>Eremanthus</i>	1,132	0,571
17	Richness	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Connected	1	No data	<i>Eucalyptus</i>	0,555	0,432
17	Richness	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Connected	4	No data	<i>Eucalyptus</i>	-1,347	0,449
17	Abundance	Dung beetle	Native	Tropical	Monoculture	Commercial	Connected	4	No data	<i>Eremanthus</i>	-9,430	8,160
17	Abundance	Dung beetle	Native	Tropical	Monoculture	Commercial	Connected	6	No data	<i>Eremanthus</i>	-13,773	19,803
17	Abundance	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Connected	1	No data	<i>Eucalyptus</i>	-5,870	3,621
17	Abundance	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Connected	4	No data	<i>Eucalyptus</i>	-9,247	6,808
17	Abundance	Dung beetle	Native	Tropical	Monoculture	Commercial	Connected	4	No data	<i>Eremanthus</i>	1,805	0,580
17	Abundance	Dung beetle	Native	Tropical	Monoculture	Commercial	Connected	6	No data	<i>Eremanthus</i>	1,416	0,611
17	Abundance	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Connected	1	No data	<i>Eucalyptus</i>	0,641	0,437
17	Abundance	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Connected	4	No data	<i>Eucalyptus</i>	0,167	0,368
18	Abundance	Ground beetle	Exotic	Non-tropical	Monoculture	No data	Isolated	No data	No data	<i>Pinus</i>	0,434	0,409
18	Abundance	Ground beetle	Exotic	Non-tropical	Monoculture	No data	Isolated	No data	No data	<i>Pinus</i>	-0,737	0,427
19	Richness	Ground beetle	Native	Non-tropical	Monoculture	No data	No data	9	No data	<i>Pinus</i>	3,498	1,686
19	Richness	Ground beetle	Native	Non-tropical	Monoculture	No data	No data	8	No data	<i>Pinus</i>	0,152	0,836
19	Richness	Ground beetle	Native	Non-tropical	Monoculture	No data	No data	9	No data	<i>Pinus</i>	5,828	3,497
19	Richness	Ground beetle	Native	Non-tropical	Monoculture	No data	No data	8	No data	<i>Pinus</i>	2,514	1,465
19	Abundance	Ground beetle	Native	Non-tropical	Monoculture	No data	No data	9	No data	<i>Pinus</i>	-0,410	0,681
19	Abundance	Ground beetle	Native	Non-tropical	Monoculture	No data	No data	8	No data	<i>Pinus</i>	-0,343	0,845
19	Abundance	Ground beetle	Native	Non-tropical	Monoculture	No data	No data	9	No data	<i>Pinus</i>	-1,700	0,908
19	Abundance	Ground beetle	Native	Non-tropical	Monoculture	No data	No data	8	No data	<i>Pinus</i>	-1,331	1,010



20	Richness	Ground beetle	Native	Non-tropical	Mixed	No data	No data	95	No data	<i>Pinus and Abies</i>	0,660	0,527
20	Richness	Ground beetle	Native	Non-tropical	Mixed	No data	No data	95	No data	<i>Pinus and Abies</i>	0,973	0,559
20	Abundance	Ground beetle	Native	Non-tropical	Mixed	No data	No data	95	No data	<i>Pinus and Abies</i>	-0,499	0,516
20	Abundance	Ground beetle	Native	Non-tropical	Mixed	No data	No data	95	No data	<i>Pinus and Abies</i>	1,045	0,568
21	Richness	Ground beetle	Native	Non-tropical	Monoculture	Protective	No data	70	No data	<i>Picea</i>	0,906	0,307
21	Richness	Ground beetle	Native	Non-tropical	Monoculture	Protective	No data	120	No data	<i>Fagus</i>	0,066	0,267
21	Richness	Ground beetle	Native	Non-tropical	Monoculture	Protective	No data	120	No data	<i>Fagus</i>	0,685	0,266
21	Richness	Ground beetle	Native	Non-tropical	Monoculture	Protective	No data	70	No data	<i>Picea</i>	0,640	0,346
21	Richness	Ground beetle	Native	Non-tropical	Monoculture	Protective	No data	120	No data	<i>Fagus</i>	0,394	0,198
21	Richness	Ground beetle	Native	Non-tropical	Monoculture	Protective	No data	120	No data	<i>Fagus</i>	0,729	0,186
21	Richness	Ground beetle	Native	Non-tropical	Monoculture	Protective	No data	70	No data	<i>Picea</i>	1,110	0,259
21	Richness	Ground beetle	Native	Non-tropical	Monoculture	Protective	No data	120	No data	<i>Fagus</i>	1,049	0,240
21	Richness	Rove beetle	Native	Non-tropical	Monoculture	Protective	No data	70	No data	<i>Picea</i>	0,150	0,284
21	Richness	Rove beetle	Native	Non-tropical	Monoculture	Protective	No data	120	No data	<i>Fagus</i>	-0,948	0,289
21	Richness	Rove beetle	Native	Non-tropical	Monoculture	Protective	No data	120	No data	<i>Fagus</i>	-0,962	0,276
21	Richness	Rove beetle	Native	Non-tropical	Monoculture	Protective	No data	70	No data	<i>Picea</i>	0,500	0,341
21	Richness	Rove beetle	Native	Non-tropical	Monoculture	Protective	No data	120	No data	<i>Fagus</i>	0,229	0,196
21	Richness	Rove beetle	Native	Non-tropical	Monoculture	Protective	No data	120	No data	<i>Fagus</i>	0,380	0,177
21	Richness	Rove beetle	Native	Non-tropical	Monoculture	Protective	No data	70	No data	<i>Picea</i>	0,787	0,242
21	Richness	Rove beetle	Native	Non-tropical	Monoculture	Protective	No data	120	No data	<i>Fagus</i>	1,436	0,263
21	Abundance	Ground beetle	Native	Non-tropical	Monoculture	Protective	No data	70	No data	<i>Picea</i>	-0,175	0,284
21	Abundance	Ground beetle	Native	Non-tropical	Monoculture	Protective	No data	120	No data	<i>Fagus</i>	-1,285	0,308
21	Abundance	Ground beetle	Native	Non-tropical	Monoculture	Protective	No data	120	No data	<i>Fagus</i>	-0,773	0,269
21	Abundance	Ground beetle	Native	Non-tropical	Monoculture	Protective	No data	70	No data	<i>Picea</i>	-0,304	0,336
21	Abundance	Ground beetle	Native	Non-tropical	Monoculture	Protective	No data	120	No data	<i>Fagus</i>	-0,553	0,202
21	Abundance	Ground beetle	Native	Non-tropical	Monoculture	Protective	No data	120	No data	<i>Fagus</i>	0,052	0,174
21	Abundance	Ground beetle	Native	Non-tropical	Monoculture	Protective	No data	70	No data	<i>Picea</i>	-0,720	0,240
21	Abundance	Ground beetle	Native	Non-tropical	Monoculture	Protective	No data	120	No data	<i>Fagus</i>	0,816	0,230
21	Abundance	Rove beetle	Native	Non-tropical	Monoculture	Protective	No data	70	No data	<i>Picea</i>	-0,250	0,285
21	Abundance	Rove beetle	Native	Non-tropical	Monoculture	Protective	No data	120	No data	<i>Fagus</i>	-1,257	0,306
21	Abundance	Rove beetle	Native	Non-tropical	Monoculture	Protective	No data	120	No data	<i>Fagus</i>	0,241	0,257
21	Abundance	Rove beetle	Native	Non-tropical	Monoculture	Protective	No data	70	No data	<i>Picea</i>	-0,818	0,354
21	Abundance	Rove beetle	Native	Non-tropical	Monoculture	Protective	No data	120	No data	<i>Fagus</i>	-0,472	0,200
21	Abundance	Rove beetle	Native	Non-tropical	Monoculture	Protective	No data	120	No data	<i>Fagus</i>	-0,199	0,175
21	Abundance	Rove beetle	Native	Non-tropical	Monoculture	Protective	No data	70	No data	<i>Picea</i>	-1,108	0,259
21	Abundance	Rove beetle	Native	Non-tropical	Monoculture	Protective	No data	120	No data	<i>Fagus</i>	0,857	0,232
22	Abundance	Ground beetle	Native	Non-tropical	Mixed	Protective	Connected	50	< 400 ha	<i>Populus</i>	-0,689	0,071
23	Richness	Rove beetle	Native	Non-tropical	Monoculture	Protective	No data	40	< 400 ha	<i>Picea</i>	0,530	0,690
23	Richness	Rove beetle	Native	Non-tropical	Mixed	Protective	No data	40	< 400 ha	<i>Larix</i>	1,659	0,896
23	Richness	Rove beetle	Native	Non-tropical	Monoculture	Protective	No data	40	< 400 ha	<i>Picea</i>	-0,682	0,705
23	Richness	Rove beetle	Native	Non-tropical	Mixed	Protective	No data	40	< 400 ha	<i>Larix</i>	-0,365	0,678
23	Richness	Rove beetle	Native	Non-tropical	Monoculture	Protective	No data	40	< 400 ha	<i>Picea</i>	-2,275	1,098
23	Richness	Rove beetle	Native	Non-tropical	Mixed	Protective	No data	40	< 400 ha	<i>Larix</i>	-4,084	2,056



23	Abundance	Rove beetle	Native	Non-tropical	Monoculture	Protective	No data	40	< 400 ha	<i>Picea</i>	3,224	1,533
23	Abundance	Rove beetle	Native	Non-tropical	Mixed	Protective	No data	40	< 400 ha	<i>Larix</i>	0,668	0,704
23	Abundance	Rove beetle	Native	Non-tropical	Monoculture	Protective	No data	40	< 400 ha	<i>Picea</i>	5,468	3,158
23	Abundance	Rove beetle	Native	Non-tropical	Mixed	Protective	No data	40	< 400 ha	<i>Larix</i>	0,612	0,698
23	Abundance	Rove beetle	Native	Non-tropical	Monoculture	Protective	No data	40	< 400 ha	<i>Picea</i>	1,881	0,961
23	Abundance	Rove beetle	Native	Non-tropical	Mixed	Protective	No data	40	< 400 ha	<i>Larix</i>	-0,056	0,667
24	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	Protective	No data	27	No data	<i>Picea</i>	-1,470	0,141
24	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	Protective	No data	27	No data	<i>Picea</i>	-2,526	0,200
24	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	Commercial	No data	34	No data	<i>Picea</i>	-0,892	0,122
24	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	Commercial	No data	34	No data	<i>Picea</i>	-0,816	0,120
24	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	Protective	No data	27	No data	<i>Picea</i>	-1,740	0,153
24	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	Protective	No data	27	No data	<i>Picea</i>	-4,624	0,408
24	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	Commercial	No data	34	No data	<i>Picea</i>	-1,288	0,134
24	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	Commercial	No data	34	No data	<i>Picea</i>	-1,106	0,128
24	Abundance	Ground beetle	Exotic	Non-tropical	Monoculture	Protective	No data	27	No data	<i>Picea</i>	-2,379	0,190
24	Abundance	Ground beetle	Exotic	Non-tropical	Monoculture	Protective	No data	27	No data	<i>Picea</i>	-2,596	0,205
24	Abundance	Ground beetle	Exotic	Non-tropical	Monoculture	Commercial	No data	34	No data	<i>Picea</i>	-0,902	0,122
24	Abundance	Ground beetle	Exotic	Non-tropical	Monoculture	Commercial	No data	34	No data	<i>Picea</i>	-0,628	0,117
24	Abundance	Ground beetle	Exotic	Non-tropical	Monoculture	Protective	No data	27	No data	<i>Picea</i>	-2,535	0,200
24	Abundance	Ground beetle	Exotic	Non-tropical	Monoculture	Protective	No data	27	No data	<i>Picea</i>	-2,905	0,228
24	Abundance	Ground beetle	Exotic	Non-tropical	Monoculture	Commercial	No data	34	No data	<i>Picea</i>	-0,910	0,123
24	Abundance	Ground beetle	Exotic	Non-tropical	Monoculture	Commercial	No data	34	No data	<i>Picea</i>	-0,350	0,113
25	Abundance	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	5	< 400 ha	<i>Picea</i>	0,892	0,137
25	Abundance	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	15	< 400 ha	<i>Picea</i>	-0,512	0,129
25	Abundance	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	30	< 400 ha	<i>Picea</i>	-0,851	0,136
25	Abundance	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	50	< 400 ha	<i>Picea</i>	-1,122	0,145
26	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	5	< 400 ha	<i>Picea</i>	-1,523	0,516
26	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	5	< 400 ha	<i>Picea</i>	-3,982	1,193
26	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	15	< 400 ha	<i>Picea</i>	-5,693	2,021
26	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	15	< 400 ha	<i>Picea</i>	-4,750	1,528
26	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	30	< 400 ha	<i>Picea</i>	-5,417	1,867
26	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	30	< 400 ha	<i>Picea</i>	-5,988	2,193
26	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	50	< 400 ha	<i>Picea</i>	-5,231	1,768
26	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	50	< 400 ha	<i>Picea</i>	-4,262	1,308
26	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	5	< 400 ha	<i>Picea</i>	-1,300	0,484
26	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	5	< 400 ha	<i>Picea</i>	-2,979	0,844
26	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	15	< 400 ha	<i>Picea</i>	-3,608	1,051
26	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	15	< 400 ha	<i>Picea</i>	-3,699	1,084
26	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	30	< 400 ha	<i>Picea</i>	-3,813	1,127
26	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	30	< 400 ha	<i>Picea</i>	-3,873	1,150
26	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	50	< 400 ha	<i>Picea</i>	-3,503	1,014
26	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	50	< 400 ha	<i>Picea</i>	-3,267	0,934
27	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	5	< 400 ha	<i>Picea</i>	-1,568	0,163
27	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	15	< 400 ha	<i>Picea</i>	-3,702	0,339
27	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	30	< 400 ha	<i>Picea</i>	-2,695	0,238
27	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	50	< 400 ha	<i>Picea</i>	-3,136	0,279

27	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	5	< 400 ha	<i>Picea</i>	-0,934	0,139
27	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	15	< 400 ha	<i>Picea</i>	-0,630	0,131
27	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	30	< 400 ha	<i>Picea</i>	-1,709	0,171
27	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	50	< 400 ha	<i>Picea</i>	-1,184	0,147
28	Richness	Ground beetle	Native	Non-tropical	Monoculture	Commercial	No data	5	< 400 ha	<i>Quercus</i>	-1,653	0,112
28	Richness	Ground beetle	Native	Non-tropical	Monoculture	Commercial	No data	15	< 400 ha	<i>Quercus</i>	-0,311	0,084
28	Richness	Ground beetle	Native	Non-tropical	Monoculture	Commercial	No data	45	< 400 ha	<i>Quercus</i>	0,786	0,090
28	Abundance	Ground beetle	Native	Non-tropical	Monoculture	Commercial	No data	5	< 400 ha	<i>Quercus</i>	-1,502	0,107
28	Abundance	Ground beetle	Native	Non-tropical	Monoculture	Commercial	No data	15	< 400 ha	<i>Quercus</i>	-0,190	0,084
28	Abundance	Ground beetle	Native	Non-tropical	Monoculture	Commercial	No data	45	< 400 ha	<i>Quercus</i>	0,967	0,093
29	Richness	Ground beetle	Exotic	Non-tropical	Mixed	Commercial	No data	No data	< 400 ha	<i>Pinus</i>	0,479	0,290
29	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	Commercial	No data	No data	< 400 ha	<i>Pinus</i>	0,343	0,141
29	Richness	Ground beetle	Exotic	Non-tropical	Mixed	Commercial	No data	No data	< 400 ha	<i>Pinus</i>	-0,421	0,317
29	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	Commercial	No data	No data	< 400 ha	<i>Pinus</i>	-0,926	0,183
29	Richness	Ground beetle	Exotic	Non-tropical	Mixed	Commercial	No data	No data	< 400 ha	<i>Pinus</i>	0,674	0,342
29	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	Commercial	No data	No data	< 400 ha	<i>Pinus</i>	0,639	0,188
29	Abundance	Ground beetle	Exotic	Non-tropical	Mixed	Commercial	No data	No data	< 400 ha	<i>Pinus</i>	0,560	0,293
29	Abundance	Ground beetle	Exotic	Non-tropical	Monoculture	Commercial	No data	No data	< 400 ha	<i>Pinus</i>	0,330	0,141
29	Abundance	Ground beetle	Exotic	Non-tropical	Mixed	Commercial	No data	No data	< 400 ha	<i>Pinus</i>	0,580	0,323
29	Abundance	Ground beetle	Exotic	Non-tropical	Monoculture	Commercial	No data	No data	< 400 ha	<i>Pinus</i>	0,344	0,169
29	Abundance	Ground beetle	Exotic	Non-tropical	Mixed	Commercial	No data	No data	< 400 ha	<i>Pinus</i>	0,994	0,363
29	Abundance	Ground beetle	Exotic	Non-tropical	Monoculture	Commercial	No data	No data	< 400 ha	<i>Pinus</i>	0,911	0,197
30	Richness	Dung beetle	Exotic	Tropical	Monoculture	Protective	Connected	No data	No data	<i>Fraxinus</i>	-1,892	0,289
30	Abundance	Dung beetle	Exotic	Tropical	Monoculture	Protective	Connected	No data	No data	<i>Fraxinus</i>	-2,127	0,313
30	Abundance	Dung beetle	Native	Tropical	Monoculture	Protective	Connected	No data	No data	<i>Alnus</i>	-1,411	0,250
30	Abundance	Dung beetle	Exotic	Tropical	Monoculture	Protective	Connected	No data	No data	<i>Fraxinus</i>	0,977	0,224
30	Abundance	Dung beetle	Native	Tropical	Monoculture	Protective	Connected	No data	No data	<i>Alnus</i>	1,373	0,247
31	Richness	Rove beetle	Native	Tropical	Monoculture	Protective	No data	5	No data	<i>Alnus</i>	-0,866	0,199
31	Richness	Rove beetle	Native	Tropical	Monoculture	Protective	No data	10	No data	<i>Alnus</i>	0,250	0,183
31	Richness	Rove beetle	Native	Tropical	Monoculture	Protective	No data	5	No data	<i>Alnus</i>	-0,239	0,183
31	Richness	Rove beetle	Native	Tropical	Monoculture	Protective	No data	10	No data	<i>Alnus</i>	1,066	0,208
31	Abundance	Rove beetle	Native	Tropical	Monoculture	Protective	No data	5	No data	<i>Alnus</i>	-0,492	0,187
31	Abundance	Rove beetle	Native	Tropical	Monoculture	Protective	No data	10	No data	<i>Alnus</i>	0,374	0,185
31	Abundance	Rove beetle	Native	Tropical	Monoculture	Protective	No data	5	No data	<i>Alnus</i>	0,565	0,189
31	Abundance	Rove beetle	Native	Tropical	Monoculture	Protective	No data	10	No data	<i>Alnus</i>	1,477	0,231
32	Richness	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Isolated	4	> 1000 ha	<i>Eucalyptus</i>	-2,383	0,570
32	Richness	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Isolated	4	> 1000 ha	<i>Eucalyptus</i>	-1,044	0,379
32	Richness	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Isolated	4	> 1000 ha	<i>Eucalyptus</i>	-1,975	0,496
32	Abundance	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Isolated	4	> 1000 ha	<i>Eucalyptus</i>	-1,749	0,461
32	Abundance	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Isolated	4	> 1000 ha	<i>Eucalyptus</i>	-0,767	0,358
32	Abundance	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Isolated	4	> 1000 ha	<i>Eucalyptus</i>	-1,077	0,382
33	Richness	Rove beetle	Native	Non-tropical	Monoculture	Protective	No data	No data	No data	<i>Quercus</i>	-2,533	0,721
33	Richness	Rove beetle	Native	Non-tropical	Monoculture	Protective	No data	31	No data	<i>Quercus</i>	-3,610	1,052
33	Richness	Rove beetle	Exotic	Non-tropical	Monoculture	Protective	No data	39	No data	<i>Pinus</i>	-4,332	1,338
33	Richness	Rove beetle	Exotic	Non-tropical	Monoculture	Protective	No data	30	No data	<i>Robinia</i>	-3,265	0,933
33	Abundance	Rove beetle	Native	Non-tropical	Monoculture	Protective	No data	No data	No data	<i>Quercus</i>	-1,813	0,564

33	Abundance	Rove beetle	Native	Non-tropical	Monoculture	Protective	No data	31	No data	<i>Quercus</i>	-4,805	1,555
33	Abundance	Rove beetle	Exotic	Non-tropical	Monoculture	Protective	No data	39	No data	<i>Pinus</i>	-2,506	0,714
33	Abundance	Rove beetle	Exotic	Non-tropical	Monoculture	Protective	No data	30	No data	<i>Robinia</i>	-2,396	0,687
34	Abundance	Dung beetle	Exotic	Tropical	Monoculture	No data	Connected	No data	No data	<i>Eucalyptus</i>	-0,735	0,214
34	Abundance	Dung beetle	Exotic	Tropical	Monoculture	No data	Isolated	No data	No data	<i>Eucalyptus</i>	-0,733	0,213
34	Abundance	Dung beetle	Exotic	Tropical	Monoculture	No data	Connected	No data	No data	<i>Eucalyptus</i>	-0,790	0,216
34	Abundance	Dung beetle	Exotic	Tropical	Monoculture	No data	Isolated	No data	No data	<i>Eucalyptus</i>	-0,783	0,215
35	Richness	Dung beetle	Exotic	Non-tropical	Monoculture	Commercial	Isolated	10	< 400 ha	<i>Robinia</i>	0,433	0,146
35	Richness	Dung beetle	Exotic	Non-tropical	Monoculture	Commercial	Isolated	10	< 400 ha	<i>Robinia</i>	-0,173	0,143
35	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	Commercial	Isolated	10	< 400 ha	<i>Robinia</i>	1,040	0,162
35	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	Commercial	Isolated	10	< 400 ha	<i>Robinia</i>	0,658	0,151
35	Abundance	Dung beetle	Exotic	Non-tropical	Monoculture	Commercial	Isolated	10	< 400 ha	<i>Robinia</i>	0,355	0,145
35	Abundance	Dung beetle	Exotic	Non-tropical	Monoculture	Commercial	Isolated	10	< 400 ha	<i>Robinia</i>	-0,517	0,148
35	Abundance	Ground beetle	Exotic	Non-tropical	Monoculture	Commercial	Isolated	10	< 400 ha	<i>Robinia</i>	1,015	0,161
35	Abundance	Ground beetle	Exotic	Non-tropical	Monoculture	Commercial	Isolated	10	< 400 ha	<i>Robinia</i>	1,666	0,192
36	Richness	Dung beetle	Exotic	Non-tropical	Monoculture	No data	Connected	No data	No data	<i>Pinus</i>	0,976	0,165
36	Richness	Dung beetle	Exotic	Non-tropical	Monoculture	No data	Connected	No data	No data	<i>Pinus</i>	0,506	0,083
36	Richness	Dung beetle	Exotic	Non-tropical	Monoculture	No data	Connected	No data	No data	<i>Pinus</i>	0,983	0,165
36	Richness	Dung beetle	Exotic	Non-tropical	Monoculture	No data	Connected	No data	No data	<i>Pinus</i>	0,515	0,083
36	Richness	Dung beetle	Exotic	Non-tropical	Monoculture	No data	Connected	No data	No data	<i>Pinus</i>	1,162	0,171
36	Richness	Dung beetle	Exotic	Non-tropical	Monoculture	No data	Connected	No data	No data	<i>Pinus</i>	0,729	0,085
36	Richness	Dung beetle	Exotic	Non-tropical	Monoculture	No data	Connected	No data	No data	<i>Pinus</i>	1,299	0,176
36	Richness	Dung beetle	Exotic	Non-tropical	Monoculture	No data	Connected	No data	No data	<i>Pinus</i>	0,788	0,086
37	Abundance	Ground beetle	Exotic	Non-tropical	Monoculture	Commercial	No data	20	No data	<i>Pinus</i>	-0,791	0,240
37	Abundance	Ground beetle	Exotic	Non-tropical	Monoculture	Commercial	No data	1	No data	<i>Pinus</i>	-1,028	0,252
37	Abundance	Ground beetle	Exotic	Non-tropical	Monoculture	Commercial	No data	1	No data	<i>Pinus</i>	-0,810	0,240
38	Richness	Rove beetle	Native	Tropical	Monoculture	Commercial	No data	30	< 400 ha	<i>Tectona</i>	-1,838	0,948
38	Richness	Rove beetle	Native	Tropical	Monoculture	Commercial	No data	30	< 400 ha	<i>Tectona</i>	0,003	0,667
38	Richness	Rove beetle	Native	Tropical	Monoculture	Commercial	No data	30	< 400 ha	<i>Tectona</i>	-3,310	1,580
38	Richness	Rove beetle	Native	Tropical	Monoculture	Commercial	No data	30	< 400 ha	<i>Tectona</i>	-2,976	1,405
38	Richness	Rove beetle	Native	Tropical	Monoculture	Commercial	No data	30	< 400 ha	<i>Tectona</i>	-2,938	1,386
38	Richness	Rove beetle	Native	Tropical	Monoculture	Commercial	No data	30	< 400 ha	<i>Tectona</i>	-0,663	0,703
38	Richness	Rove beetle	Native	Tropical	Monoculture	Commercial	No data	30	< 400 ha	<i>Tectona</i>	-3,887	1,926
38	Richness	Rove beetle	Native	Tropical	Monoculture	Commercial	No data	30	< 400 ha	<i>Tectona</i>	-1,103	0,768
38	Abundance	Rove beetle	Native	Tropical	Monoculture	Commercial	No data	30	< 400 ha	<i>Tectona</i>	-2,256	1,091
38	Abundance	Rove beetle	Native	Tropical	Monoculture	Commercial	No data	30	< 400 ha	<i>Tectona</i>	-0,149	0,669
38	Abundance	Rove beetle	Native	Tropical	Monoculture	Commercial	No data	30	< 400 ha	<i>Tectona</i>	-3,025	1,429
38	Abundance	Rove beetle	Native	Tropical	Monoculture	Commercial	No data	30	< 400 ha	<i>Tectona</i>	-2,117	1,040
38	Abundance	Rove beetle	Native	Tropical	Monoculture	Commercial	No data	30	< 400 ha	<i>Tectona</i>	-3,060	1,447
38	Abundance	Rove beetle	Native	Tropical	Monoculture	Commercial	No data	30	< 400 ha	<i>Tectona</i>	-1,329	0,814
38	Abundance	Rove beetle	Native	Tropical	Monoculture	Commercial	No data	30	< 400 ha	<i>Tectona</i>	-3,197	1,518
38	Abundance	Rove beetle	Native	Tropical	Monoculture	Commercial	No data	30	< 400 ha	<i>Tectona</i>	-1,585	0,876
39	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	Commercial	Isolated	13,5	> 1000 ha	<i>Pinus</i>	-2,157	0,527
39	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	Commercial	Isolated	13,5	> 1000 ha	<i>Pinus</i>	-1,410	0,416
39	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	Commercial	Isolated	13,5	> 1000 ha	<i>Pinus</i>	-0,759	0,357
39	Abundance	Ground beetle	Exotic	Non-tropical	Monoculture	Commercial	Isolated	13,5	> 1000 ha	<i>Pinus</i>	-2,053	0,509

39	Abundance	Ground beetle	Exotic	Non-tropical	Monoculture	Commercial	Isolated	13,5	> 1000 ha	<i>Pinus</i>	-1,607	0,441
39	Abundance	Ground beetle	Exotic	Non-tropical	Monoculture	Commercial	Isolated	13,5	> 1000 ha	<i>Pinus</i>	-0,683	0,353
40	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	Commercial	No data	2,5	< 400 ha	<i>Pinus</i>	1,610	0,530
40	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	Commercial	No data	11	< 400 ha	<i>Pinus</i>	-1,142	0,465
40	Richness	Ground beetle	Exotic	Non-tropical	Mixed	Commercial	No data	42,5	< 400 ha	<i>Pinus, Fagus, Fraxinus, Quercus</i>	-2,740	0,775
40	Richness	Ground beetle	Exotic	Non-tropical	Mixed	Commercial	No data	80	< 400 ha	<i>Pinus, Fraxinus, Ilex, Quercus</i>	-0,956	0,446
40	Abundance	Ground beetle	Exotic	Non-tropical	Monoculture	Commercial	No data	2,5	< 400 ha	<i>Pinus</i>	1,246	0,478
40	Abundance	Ground beetle	Exotic	Non-tropical	Monoculture	Commercial	No data	11	< 400 ha	<i>Pinus</i>	-1,081	0,458
40	Abundance	Ground beetle	Exotic	Non-tropical	Mixed	Commercial	No data	42,5	< 400 ha	<i>Pinus, Fagus, Fraxinus, Quercus</i>	-1,027	0,453
40	Abundance	Ground beetle	Exotic	Non-tropical	Mixed	Commercial	No data	80	< 400 ha	<i>Pinus, Fraxinus, Ilex, Quercus</i>	0,811	0,433
41	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	10	< 400 ha	<i>Pinus</i>	-2,963	0,839
41	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	40	< 400 ha	<i>Pinus</i>	-4,048	1,219
41	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	80	< 400 ha	<i>Pinus</i>	-2,990	0,847
41	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	10	< 400 ha	<i>Pinus</i>	-1,522	0,516
41	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	40	< 400 ha	<i>Pinus</i>	-2,596	0,737
41	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	80	< 400 ha	<i>Pinus</i>	-1,359	0,492
41	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	10	< 400 ha	<i>Pinus</i>	-1,161	0,467
41	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	40	< 400 ha	<i>Pinus</i>	-2,563	0,728
41	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	80	< 400 ha	<i>Pinus</i>	-0,926	0,443
41	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	10	< 400 ha	<i>Pinus</i>	-1,201	0,472
41	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	40	< 400 ha	<i>Pinus</i>	-2,740	0,775
41	Richness	Ground beetle	Exotic	Non-tropical	Monoculture	No data	No data	80	< 400 ha	<i>Pinus</i>	-0,984	0,448
42	Abundance	Ground beetle	Exotic	Tropical	Monoculture	No data	No data	14	No data	<i>Acacia</i>	0,000	0,167
42	Abundance	Rove beetle	Exotic	Tropical	Monoculture	No data	No data	14	No data	<i>Acacia</i>	-0,304	0,169
43	Abundance	Dung beetle	Exotic	Tropical	Monoculture	Commercial	Connected	7,5	< 400 ha	<i>Acacia</i>	-3,835	0,875
44	Abundance	Ground beetle	Native	Non-tropical	Monoculture	Protective	No data	No data	No data	<i>Pinus</i>	-0,320	0,506
44	Abundance	Ground beetle	Native	Non-tropical	Monoculture	Protective	No data	No data	No data	<i>Larix</i>	0,631	0,525
44	Abundance	Ground beetle	Native	Non-tropical	Monoculture	Protective	No data	No data	No data	<i>Pinus</i>	-0,282	0,505
44	Abundance	Ground beetle	Native	Non-tropical	Monoculture	Protective	No data	No data	No data	<i>Larix</i>	1,027	0,566
44	Abundance	Ground beetle	Native	Non-tropical	Monoculture	Protective	No data	No data	No data	<i>Pinus</i>	-1,756	0,693
44	Abundance	Ground beetle	Native	Non-tropical	Monoculture	Protective	No data	No data	No data	<i>Larix</i>	-0,435	0,512
45	Abundance	Ground beetle	Native	Non-tropical	Monoculture	No data	No data	30	< 400 ha	<i>Larix</i>	1,122	0,042
45	Abundance	Ground beetle	Native	Non-tropical	Monoculture	No data	No data	30	< 400 ha	<i>Larix</i>	-0,199	0,037
46	Richness	Ground beetle	Native	Non-tropical	Monoculture	No data	No data	5	No data	<i>Larix</i>	-6,800	3,019
46	Richness	Ground beetle	Native	Non-tropical	Monoculture	No data	No data	15	No data	<i>Larix</i>	-3,464	1,117
46	Richness	Ground beetle	Native	Non-tropical	Monoculture	No data	No data	45	No data	<i>Larix</i>	-6,457	2,766
46	Abundance	Ground beetle	Native	Non-tropical	Monoculture	No data	No data	5	No data	<i>Larix</i>	-7,560	3,625



46	Abundance	Ground beetle	Native	Non-tropical	Monoculture	No data	No data	15	No data	<i>Larix</i>	-5,280	1,999
46	Abundance	Ground beetle	Native	Non-tropical	Monoculture	No data	No data	45	No data	<i>Larix</i>	-5,419	2,082
47	Richness	Ground beetle	Native	Non-tropical	Monoculture	No data	No data	44	< 400 ha	<i>Picea</i>	1,269	0,801
47	Richness	Ground beetle	Native	Non-tropical	Mixed	No data	No data	44	< 400 ha	<i>Larix</i>	2,388	1,142
47	Richness	Ground beetle	Native	Non-tropical	Monoculture	No data	No data	44	< 400 ha	<i>Picea</i>	0,667	0,704
47	Richness	Ground beetle	Native	Non-tropical	Mixed	No data	No data	44	< 400 ha	<i>Larix</i>	2,054	1,018
47	Richness	Ground beetle	Native	Non-tropical	Monoculture	No data	No data	44	< 400 ha	<i>Picea</i>	-2,912	1,373
47	Richness	Ground beetle	Native	Non-tropical	Mixed	No data	No data	44	< 400 ha	<i>Larix</i>	-0,729	0,711
47	Abundance	Ground beetle	Native	Non-tropical	Monoculture	No data	No data	44	< 400 ha	<i>Picea</i>	2,047	1,016
47	Abundance	Ground beetle	Native	Non-tropical	Mixed	No data	No data	44	< 400 ha	<i>Larix</i>	2,134	1,046
47	Abundance	Ground beetle	Native	Non-tropical	Monoculture	No data	No data	44	< 400 ha	<i>Picea</i>	1,816	0,941
47	Abundance	Ground beetle	Native	Non-tropical	Mixed	No data	No data	44	< 400 ha	<i>Larix</i>	1,767	0,927
47	Abundance	Ground beetle	Native	Non-tropical	Monoculture	No data	No data	44	< 400 ha	<i>Picea</i>	0,392	0,679
47	Abundance	Ground beetle	Native	Non-tropical	Mixed	No data	No data	44	< 400 ha	<i>Larix</i>	-0,221	0,671
48	Abundance	Ground beetle	Exotic	Non-tropical	Monoculture	Commercial	No data	6,5	No data	<i>Eucalyptus</i>	-1,780	0,233
48	Abundance	Ground beetle	Exotic	Non-tropical	Monoculture	Commercial	No data	6,5	No data	<i>Eucalyptus</i>	-1,884	0,241

**References:** 1 = Audino *et al.* (2014) *Biological Conservation*, **169**, 248-257; 2 = Bains *et al.* (2012) *Forest Ecology and Management*, **286**, 183-191; 3 = Balog *et al.* (2012) *North-Western Journal of Zoology*, **8**, 215-222; 4 = Beiroz *et al.* (2016) *Insect Conservation and Diversity*, **10**, 94-106; 5 = Berndt *et al.* (2008) *Biodiversity and Conservation*, **17**, 1171-1185; 6 = Bowie *et al.* (2019) *New Zealand Journal of Ecology*, **43**, 1-11; 7 = Cheng *et al.* (2018) *Acta Ecologica Sinica*, **38**, 3097-3019; 8 = da Silva *et al.* (2008) *Agriculture, Ecosystems and Environment*, **124**, 270-274; 9 = Davis *et al.* (2001) *Journal of Applied Ecology*, **38**, 593-616; 10 = Derhé *et al.* (2016) *Journal of Applied Ecology*, **53**, 1714-1724; 11 = Fally & Gormally (1998) *Forest Ecology and Management*, **110**, 263-273; 12 = Gardner *et al.* (2008) *Journal of Applied Ecology*, **45**, 883-893; 13 = Giménez-Gómez *et al.* (2018a) *Insect Conservation and Diversity*, **11**, 554-564; 14 = Giménez-Gómez *et al.* (2018b) *Biodiversity and Conservation*, **27**, 3201-3213; 15 = Gollan *et al.* (2011) *Insect Conservation and Diversity*, **4**, 123-131; 16 = González-Vainer *et al.* (2012) *Neotropical Entomology*, **41**, 366-374; 17 = Gries *et al.* (2012) *Insect Conservation and Diversity*, **5**, 175-185; 18 = Henríquez *et al.* (2009) *Acta Oecologica*, **35**, 811-818; 19 = Ings & Hartley (1999) *Forest Ecology and Management*, **119**, 123-136; 20 = Kosewska *et al.* (2018) *Community Ecology*, **19**, 156-167; 21 = Lange *et al.* (2014) *Forest Ecology and Management*, **329**, 166-176; 22 = Lik (2010) *Annales de la Société Entomologique de France*, **46**, 425-438; 23 = Luo *et al.* (2013) *Environmental Entomology*, **42**, 7-16; 24 = Magura *et al.* (2000) *Biological Conservation*, **93**, 95-102; 25 = Magura *et al.* (2002) *European Journal of Soil Biology*, **38**, 291-295; 26 = Magura *et al.* (2003) *Biodiversity and Conservation*, **12**, 73-85; 27 = Magura *et al.* (2006) *Community Ecology*, **7**, 1-12; 28 = Magura *et al.* (2015) *Forest Ecology and Management*, **339**, 117-126; 29 = Martínez *et al.* (2009) *Annals of Forest Science*, **66**, 304-304; 30

= Medina *et al.* (2002) *Biotropica*, **34**, 181-187; 31 = Méndez-Rojas *et al.* (2012) *Revista Colombiana de Entomología*, **38**, 141-147; 32 = Milheiras *et al.* (2020) *PeerJ*, **8**, e8486; 33 = Nagy *et al.* (2015) *Journal of Insect Conservation*, **19**, 1075-1087; 34 = Niero & Hernández (2017) *Biotemas*, **30**, 37-48; 35 = Plexida *et al.* (2014) *Applied Ecology and Environmental Research*, **12**, 661-679; 36 = Pryke *et al.* (2013) *Biodiversity and Conservation*, **22**, 2857-2873; 37 = Russek *et al.* (2017) *Journal of Insect Conservation*, **21**, 943-950; 38 = Sakchoowong *et al.* (2008) *Entomological Science*, **11**, 301-313; 39 = Sweaney *et al.* (2015) *Biological Conservation*, **186**, 1-11; 40 = Taboada *et al.* (2008) *Basic and Applied Ecology*, **9**, 161-171; 41 = Taboada *et al.* (2010) *European Journal of Forestry Research*, **129**, 31-45; 42 = Tsukamoto & Sabang (2005) *Pedobiologia*, **49**, 69-80; 43 = Ueda *et al.* (2015) *Journal of Insect Conservation*, **19**, 765-780; 44 = Warrer-Thomas *et al.* (2014) *Forest Ecology and Management*, **334**, 369-376; 45 = Yu *et al.* (2004) *Entomologica Fennica*, **15**, 129-137; 46 = Yu *et al.* (2006) *Forest Ecology and Management*, **231**, 169-177; 47 = Yu *et al.* (2008) *Forest Ecology and Management*, **255**, 2617-2625; 48 = Zahn *et al.* (2009) *Applied Ecology and Environmental Research*, **7**, 297-301.

**Table S2.** Number of articles associated with forestry plantations by country

Country	Articles number
Argentina	2
Australia	3
Brazil	6
Chile	2
China	6
Colombia	2
Germany	1
Grecia	1
Hungria	2
Indonesia	2
Ireland	1
Italy	1
Malaysia	1
Mexico	1
New Zeland	2
Poland	6
Portugal	2
Rumania	1
Serbia	1
Scotland	1
South Africa	1
Spain	3
Thailand	1
Uruguay	1

**Table S3.** Observed and adjusted overall effect size values after Duval & Tweedie (2000) trim and fill procedure.

(a) Species richness				
	Studies trimmed	Effect estimate	Lower limit	Upper limit
Observed values		-1.090	-1.321	-0.859
Adjusted values	26	-0.659	-0,895	-0.422
(b) Abundance				
Observed values		-0.487	-0.703	-0.271

Adjusted values	23	-0.092	-0.326	0.142
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For Review Only



Bogotá, 30 June 2021

Dear Prof. Raphael Didham  
Editor in Chief  
Insect Conservation and Diversity

Receive a cordial greeting.

We are resubmitting the manuscript entitled “**What level of native beetle diversity can be supported by forestry plantations? A global synthesis**” (ICDIV-21-0097), by Pablo A. López-Bedoya, Tibor Magura, Felicity A. Edwards, David P. Edwards, José M. Rey-Benayas, Gábor L. Lövei & Jorge Ari Noriega, to be considered for publication as a research article in Insect Conservation and Diversity.

At the end of this letter, you will find the original correspondence with the comments of the editor and two reviewers, together with our replies. We believe these reviews have improved considerably both the quality and soundness of the research.

Thank you for considering this manuscript for publication.

Yours sincerely,

Jorge Ari Noriega, on behalf of all co-authors

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03-Jun-2021

Dear Dr. Noriega:

Manuscript ID ICDIV-21-0097 entitled "**What level of native beetle diversity can be supported by forestry plantations? A global synthesis**" which you submitted to Insect Conservation and Diversity, has been reviewed. The comments of the reviewers are included at the bottom of this letter.

The reviewers have recommended publication, but also suggest some minor to moderate revisions to your manuscript. In particular, please address the issue of whether the coverage of the meta-analysis is truly 'global' in extent, as mentioned by reviewer 1. I invite you to respond to the reviewers' comments and revise your manuscript accordingly. In your cover letter, please include a detailed point-by-point response to the reviewer comments, indicating how they have been addressed in the revised manuscript.

**R./** Thanks for your valuable comments and time. Following your advice, we reviewed the minor to moderate suggestions to the manuscript carefully, and response point-by-point to each of the comments of the two reviewers. We also address carefully the coverage issue of the meta-analysis that reviewer 1 mentioned.

We recognise that the impact of the COVID-19 pandemic may affect your ability to return your revised manuscript to us within the requested timeframe. If this is the case, please let us know.

You will be unable to make your revisions on the originally submitted version of the manuscript. Instead, revise your manuscript using a word processing program and save it on your computer. Please also highlight the changes to your manuscript within the document by using the track changes mode in MS Word or by using bold or colored text. If you do decide to use track changes then please also submit an additional version with track changes accepted for ease of reading for referees and editors.

Once the revised manuscript is prepared, you can upload it and submit it through your Author Center.

When submitting your revised manuscript, you will be able to respond to the comments made by the reviewer(s) in the space provided. You can use this space to document any changes you make to the original manuscript. In order to expedite the processing of the revised manuscript, please be as specific as possible in your response to the reviewer(s).

Because we are trying to facilitate timely publication of manuscripts submitted to Insect Conservation and Diversity, your revised manuscript should be uploaded as soon as possible. If it is not possible for you to submit your revision in a reasonable amount of time (preferably within 30 days), we may have to consider your paper as a new submission.

Once again, thank you for submitting your manuscript to Insect Conservation and Diversity and I look forward to receiving your revision.

Sincerely,

Prof. Raphael Didham

Editor, Insect Conservation and Diversity

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Associate Editor Comments to Author:

Associate Editor

Comments to the Author:

The manuscript has now been assessed by 2 reviewers. Both judgements were very positive. Nevertheless, they both made suggestions that would improve the manuscript even further. Therefore, I recommend revising the manuscript accordingly.

**R./ Thanks for your valuable comments and positive feedback. Following your advice, we reviewed the manuscript, and response to each of the comments and suggestions of the two reviewers.**

Reviewer(s)' Comments to Author:

### Reviewer 1

This paper addresses an important issue through a simple but highly relevant question in the title: “What level of native beetle diversity can be supported by forestry plantations?” It uses a meta-analysis approach to show that forestry plantations negatively affect richness and abundance of ground, rove and dung beetles compared to natural forests. Authors also show that the negative impact was most severe in plantations with exotic tree species and located in tropical biomes. I made some suggestions to improve the language but a native English speaker could help to make it more completely.

I was surprised that the boreal zone was completely absent from the paper as there is no study in North America, as well as in Scandinavian countries and Russia. This is strange as authors refer to some papers from boreal zone (Pohl from Canada, Koivula and Niemela from Finland). This is unfortunate as plantations are heavily used in these countries as recognized by authors at line 63. It would worth knowing if the main conclusions of the present meta-analysis (using small plantations, close to natural forests and using native trees) are applicable in the boreal zone. At line 64 of the introduction, authors say that there are numerous plantations in North America. Why are there none in your meta-analysis? Besides, at line 77, they report that studies found biodiversity levels within plantations to match those found in natural forests, particularly in some Nearctic and Palearctic non-tropical biomes, which in fact include the boreal zone. Rather than ignoring the boreal zone, authors should explain why no studies qualified for their meta-analysis and in the Discussion refer to the boreal zone and how their results may apply or not.

**R./** Thanks to the reviewer to point out this important issue. In our meta-analysis, articles located in boreal biomes were not included as no studies were found that met the established criteria. Despite this, different investigations in boreal systems were mentioned in the introductory section as the reviewer mentioned. When searching for scientific literature in the databases consulted, articles were obtained for the boreal biome, however, the inclusion criteria that we used filtered those articles (*e.g.*, Yi & Moldenke 2005 – USA, Pohl *et al.* 2007 – Canada, Oxbrough *et al.* 2010 – Ireland). The main two reasons of this exclusion were: i) the studies evaluate a forestry plantation without having a natural forest as a control and ii) the studies did not present the statistical dispersion measures needed to be included in the article. In some cases, these articles in boreal biomes, were limited to presenting the absolute beetle richness or abundance and did not present data on the mean or standard deviation. However, we truly believe that our findings associated with the management used in forestry plantations can be applied in different biomes (tropical and non-tropical) even without having articles from some of these specific biomes. In addition, following the advice of the reviewer, we include an explanation related to the absence of articles included from these boreal biomes in the results section (L211-215). Finally, we also include in the discussion section the possible application of our results on those boreal biomes (L406-410).

For these reasons, authors should not present their study as a “global” meta-analysis as it mainly addresses Europe and South America, with some insight into southeast Asia and Australia-New Zealand but not North America, Africa (only 1 site), Scandinavia, Russia, Japan... where plantations are also used.

**R./** We understand the point of view of the reviewer; however, we believe that the meta-analysis is a global study. We include 48 articles (12 Tropical and 36 Non-Tropical biomes) distributed in 24 countries on 5 continents. This is a global perspective without any doubt. Not including boreal biome studies, as they do not meet the inclusion criteria, is not an intentional decision and we believe is not affecting the clear global pattern that we are founding. And although countries such as the USA, Canada, Russia, and Scandinavian countries are very important, their non-inclusion does not condition that the study has a global character.

Thus, they should explain why they haven't been able to cover more countries or the boreal zone as well as other countries. Also, as they incorporated ground, rove and dung beetles into their meta-analysis, I guess that pitfall traps were used in most studies. Such information should be provided. Follows are some specific comments on various parts of the manuscript:

**R./** Done. We provide information about the absence of included studies of some non-tropical biomes (see above). Also, following the suggestion of the reviewer, we include information in the materials and methods section about the capture method used in most of the articles included (L132-133).

Abstract

L7: change “although often via the conversion of natural forest” by “often leading to the conversion of natural to artificial forests.”

**R./** Done. We made the suggested change (L7).

L10: change wording for “globally” as explained above.

**R./** Done. We delete this word in the sentence.

L20: Technically, it might not be the age of plantation that influence beetle richness and abundance but attributes that change with age (ex: coarse woody debris, forest structure, litter...) and as a result it would be better to write “Species richness and abundance of beetles significantly increased with plantation age in native plantations, but decreased in

exotic ones. Also, small plantations close to native forest had higher beetle species richness and abundance than large ones located far away from native forest.”

**R./ Done.** We made the suggested change (L20-23).

L23: change “allowing them to mature is critical to creating biodiversity-friendly forestry plantations...” by “lengthening rotations is critical for allowing biodiversity recovery in forestry plantations...”

**R./ Done.** We made the suggested change (L24-25).

## Introduction

L55: You present a meta-analysis aimed to respond to a question and it is not a good idea to respond to this question in the first sentence of the paper. I suggest removing this sentence which is also incorrect as the loss and fragmentation of natural forests is not related to timber plantation expansion but rather to enhance timber production as explained in the second sentence.

**R./ Done.** We removed this sentence of the introduction section.

L86-89: “In general, these studies show that forestry plantations composed of native or mixed species, small size, and if planted for conservation purposes can positively affect species richness or abundance of forest-inhabiting invertebrates and/or vertebrates”. Again, I suggest removing this sentence as you do not need to respond to the title question in the Introduction. Let the data speak in the Results section and keep this sentence for the Discussion.

**R./ Done.** We removed this sentence of the introduction section.

L104: change “moderate” for “modify”

**R./ Done.** We made the suggested change (L99).

L110: worldwide? I suggest rewording for reasons explained above.

**R./ Done.** We removed “worldwide” in the sentence.

L112-115: the rationale behind hypotheses should be explained. For instance, in (H1), you could remind that plantations simplify forest composition and structure and as a result, we

may expect that beetle diversity should follow the same trend. For (H2), you should explain that exotic planted trees bring natural forests farther than native trees and as a result, we may expect that a certain number of species, closely linked with conditions that developed over long time with natural forests, should suffer thus reducing diversity.

**R./ Done.** We made the suggested change (L107-109).

## Material and methods

L217: native rather than exotic species?

**R./ Done.** We change the phrase to explain that abandoned plantations of exotic species were used in conservation plans (L223).

L221: I don't understand that you have used "regenerat\*", "restor\*" or "land-use" as research terms as they are not relevant to the question asked in the title. Also, why were the terms abundance and species richness not crossed with Forest\* and Plantat\*? It would have avoided to gather 3675 articles, from which only 48 were retained, i.e. just slightly more than 1%!

**R./** Thanks to the reviewer to point out this issue. We use these search terms in order to increase the possibilities of inclusion. Different articles assess a broad spectrum of land cover or land uses (including forestry plantations) and therefore in the title, summary or keywords section do not directly refer to forestry plantations, using more general terms. In other cases, they refer to forestry plantations as artificial regeneration or restoration, so we conclude that the use of these terms within our search, can offer a greater number of results by minimizing the exclusion of results. We understand it was a great work but we are confident with our results. Following the comment of the reviewer, we include a summary explanation in methods section (L120-122).

L124: why 85% of the studies were rejected at the first step? It is not clear.

**R./ Done.** We included in the text a clear reason of rejected studies in the first step (L123-124).

L129: change "treatments stand" by "treatment stands"

**R./ Done.** We made the suggested change (L130).

## Results

Authors mainly compared the treatments between them using Qb but this test compares the homogeneity in treatments groups. In my understanding, a non-significant Qb test indicates that a significant difference between treatments is not detected because of heterogeneity of variance in compared groups. However, often there are significant differences from neutral effects for one treatment, but not for the other and these differences should also be highlighted. For instance, beetle abundance did not differ between isolated and connected plantations but the negative effect was significant for isolated plantations while it was not for connected plantations.

**R./ Done.** We agree with the reviewer and we include in the results section a more detailed associated with statistical results (L257-260, 266-267). Also see below each point.

L209: this is not apparent on the map (Fig. 1) that most studies came from Brazil, China and Poland. Maybe you will need to add a table to make it clearer.

**R./ Done.** We included an additional Table S2 in the Supplementary material (L211-212).

L226: rove beetles were also significantly less abundant than the neutral effect while abundance of ground beetles did not differ from the neutral effect.

**R./ Done.** Thanks to the reviewer to help us to correct this. We change this sentence according with the statistical results (L231).

L243-245: should remain on significant effects. Only 4 tree species produced significant negative effects on both species richness or abundance and one (*Larix*) had positive effect on abundance. You do not need to compare with *Alnus* and others as there was no significant effect on these. Also, you should compare the same tree species or group of tree species for abundance and species richness. For instance, *Acacia* and *Swietenia* are combined for abundance but not for species richness. It is important to maintain coherence in the treatments, otherwise, it raises more questions than responses.

**R./** We understand the main concern of the reviewer. In order to maintain consistency in the treatments, some of the implemented tree genera (those treatments with more than one genus) were eliminated (*Pinus* and *Abies* and *Swietenia* and *Acacia*). However, we clarify that it is very difficult to maintain coherence between the genera of trees evaluated for the beetle species richness and abundance metrics, given the different attributes of local forestry plantations. For this reason, we consider maintaining the different arboreal genera although these genera are not present for both beetle species richness and abundance (see Fig. 1, 2).

L250: I would rather write: “Likewise, the purpose (conservation or timber supply) for establishing a plantation had no effect on either beetle species richness or abundance”.

**R./ Done.** We made the suggested change (L257).

L252: for beetle abundance, the negative effect was not significantly different than the neutral effect in plantations used for conservation.

**R./ Done.** We included this suggested result in the text (L259-260).

L258: Ok, but beetle abundance was significantly lower than the neutral effect in isolated plantations while not in connected ones. This should also be highlighted.

**R./ Done.** We included this suggested result in the text (L266-267).

L263:  $r^2$  of these relationships should be provided

**R./** For meta-analyses using random effects models in the meta-regression,  $r^2$  values are not used. In this case  $\text{Tau}^2$  values are included (L276-279).

## Discussion

L288: it is not really a global meta-analysis as there is no studies coming from the boreal zone (North America, Scandinavian countries, Russia) where plantations are also widely used, as well as from Africa where there is only one study. In fact, the study covers Europe, South America, Australia-New Zealand and southeast Asia.

**R./** We already discuss this point and argue our point of view. We believe that the meta-analysis that we performed is a global study because it includes 48 articles (12 Tropical and 36 Non-Tropical biomes) distributed in 24 countries on 5 continents. Not including boreal biome studies, as they do not meet the inclusion criteria, is not an intentional decision and we believe is not affecting the global pattern that we are founding.

To simplify, I would combine the first two sentences as follows:

“Our meta-analysis on the effect of forestry plantations on native beetle diversity support our research hypotheses, revealing that species richness and abundance of ground, rove and dung beetles were generally lower than in natural forests (H1), and that location and management affected the conservation value of plantations (H2)”.

**R./ Done.** We made the suggested change (L298-301).



L297: Decline refers to time. I would rather re-write the sentence for: “Our results showed lower species richness and abundance of beetles in forestry plantations compared to natural forests”.

**R./ Done.** We made the suggested change (L306).

L303: change “... transformation of natural ecosystems into silvicultural land...” by “...conversion of natural ecosystems into intensively managed lands...”

**R./ Done.** We made the suggested change (L312-313).

L307: it is strange to refer to North American and Scandinavian studies to say that ground and rove beetles are extremely sensitive to changes in environmental and habitat characteristics caused by plantation establishment while there is no studies from these areas of the world in this meta-analysis.

**R./** We clarified this doubt above. We refer to these articles in the text, but they were not included in the analysis because they do not meet the inclusion criteria: i) the studies evaluate a forestry plantation without having a natural forest as a control and ii) the studies did not present the statistical dispersion measures needed to be included in the article.

L329: you should explain why forest specialists in tropical forests are more sensitive to habitat alteration than in non-tropical forests. Are forest specialists more prevalent in tropical forests due to the greater number of microhabitats?

**R./ Done.** We include this explanation in the text (L338-340).

L339: yes but mainly in non-tropical forests. Getting deeper in the explanation vs the previous comment could help better understand the mechanisms behind these differences and would help generalize recommendations. I think that the arguments are present in the following sentence. It just need to be widened to tropical forests and placed in appropriate place in the discussion.

**R./ Done.** We include this explanation in the text (L341-342).

L349: change “beetles reveal a marked decrease in species richness or abundance...” for “beetle richness or abundance decreased markedly...”

**R./ Done.** We made the suggested change (L359).

L394: we are not “creating” exotic forestry plantations... you should rather say: “Thus, establishing exotic forestry plantations...”

**R./ Done.** We made the suggested change (L404).

## Reviewer 2

### Comments to the Author

The authors present the results from a meta-analysis of 48 studies to compare species richness and abundance of 3 coleopteran groups between forestry plantations and natural forests. The methodology is clearly presented and consistent. The results provide a clear picture of the effects of forestry plantations when replacing « natural » forests. These results are briefly discussed and support some general recommendation for forest managers.

However, I have three minor concerns:

- The expression « natural forest » need to be defined as some forest controls used in several studies (Paillet et al, 2010 or Lang et al, 2014 for example) are actually forests that have been abandoned for only few decades. The authors also use the expression « native forest » (L 311) or « native stands » (L 386) as synonymous. Please use consistent expression throughout the manuscript.

**R./ Thanks to the reviewer to point out this important issue. We completely agree. We accepted this suggestion and reviewed the manuscript and established a consistent vocabulary (e.g., L327). Also, we included a definition of natural forest for clarity in the manuscript (L137).**

- On statement in the abstract (LL 21-22) is equivocal. As in all cases, the plantations have negative effects, one should talk about « less worse case scenario » rather than mentioning «positive effect ». Which is in line with the discussion and conclusion of the paper (« less severe negative effect », LL 347-348, or « reduce adverse effects », LL 395-396). I also recommend to take up the statement of LL 293-294 in the abstract.

**R./ Done.** We agree with this valid comment and we made the suggested change in the text (L20-25).

- The authors quoted three papers to support the idea that biodiversity levels in plantation match those found in natural forests (LL76-78). But Kerr (1999)'s and Pawson et al (2013)'s papers did not provide any data about this idea. In the latter, the authors even wrote that « the establishment of plantation forests that replace natural vegetation typically causes biodiversity losses locally ».

**R./** Done. Thanks to the reviewer to point out this issue. We have reviewed the articles in detail, and we have decided to change them to others papers that better support the idea expressed in the text (L76).

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